

# Effects of elevated carbon dioxide and increased temperature on methane and nitrous oxide fluxes: evidence from field experiments

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Climate change could alter terrestrial ecosystems, which are important sources and sinks of the potent greenhouse gases (GHGs) nitrous oxide (N<sub>2</sub>O) and methane (CH<sub>4</sub>), in ways that either stimulate or decrease the magnitude and duration of global warming. Using manipulative field experiments, we assessed how N<sub>2</sub>O and CH<sub>4</sub> soil fluxes responded to a rise in atmospheric carbon dioxide (CO<sub>2</sub>) concentration and to increased air temperature. Nitrous oxide and CH<sub>4</sub> responses varied greatly among studied ecosystems. Elevated CO<sub>2</sub> often stimulated N<sub>2</sub>O emissions in fertilized systems and CH<sub>4</sub> emissions in wetlands, peatlands, and rice paddy fields; both effects were stronger in clayey soils than in sandy upland soils. Elevated temperature, however, impacted N<sub>2</sub>O and CH<sub>4</sub> emissions inconsistently. Thus, the effects of elevated CO<sub>2</sub> concentrations on N<sub>2</sub>O and CH<sub>4</sub> emissions may further enhance global warming, but it remains unclear whether increased temperature generates positive or negative feedbacks on these GHGs in terrestrial ecosystems.

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Global warming is caused by increased atmospheric concentrations of the greenhouse gases (GHGs) carbon dioxide (CO<sub>2</sub>), nitrous oxide (N<sub>2</sub>O), and methane (CH<sub>4</sub>). Terrestrial ecosystems are important sources and sinks of these GHGs, all of which are produced and consumed through biological processes including photosynthesis, decomposition, nitrification, denitrification, methanogenesis, and CH<sub>4</sub> oxidation (eg Schlesinger 1997). Increased atmospheric CO<sub>2</sub> concentration and ele-

vated air/soil temperatures (hereafter elevated CO<sub>2</sub> and temperature; please note also that, unless stated otherwise, the following text refers to terrestrial ecosystems) can directly and indirectly alter these processes. Depending on the direction and magnitude of the alteration, elevated CO<sub>2</sub> and temperature could either accelerate or decelerate the rate of global warming.

The effects of elevated CO<sub>2</sub> and temperature on N<sub>2</sub>O and CH<sub>4</sub> fluxes in terrestrial ecosystems have been studied less frequently than the effects on CO<sub>2</sub> exchange. This is not surprising given that CO<sub>2</sub> exchange rates are usually orders of magnitude greater than the exchange rates of N<sub>2</sub>O and CH<sub>4</sub> (Schlesinger 1997). However, N<sub>2</sub>O and CH<sub>4</sub> have higher global warming potentials (GWPs) than that of CO<sub>2</sub>. Thus, although CO<sub>2</sub> is – per molecule – the most important GHG, N<sub>2</sub>O and CH<sub>4</sub> are more efficient in warming the atmosphere (the GWPs of N<sub>2</sub>O and CH<sub>4</sub> are 298 and 25 times that of CO<sub>2</sub>, respectively, over a 100-year period; Forster *et al.* 2007). Global warming is therefore more sensitive to changes in the exchange of N<sub>2</sub>O and CH<sub>4</sub> relative to that of CO<sub>2</sub>.

Process-based ecosystem models applied at regional and continental scales have recently estimated that net N<sub>2</sub>O and CH<sub>4</sub> emissions increased during the past 40 years and could further increase in the future because of elevated CO<sub>2</sub> and temperatures (Xu *et al.* 2010, 2012; Tian *et al.* 2012). Although important for long-term and large-scale predictions of climate-change feedbacks, modeling efforts still leave a lot of uncertainty, mostly due to our limited understanding of the underlying mechanisms governing N<sub>2</sub>O and CH<sub>4</sub> fluxes in different ecosystems (Tian *et al.* 2012).

## In a nutshell:

- Net emissions of nitrous oxide (N<sub>2</sub>O) and methane (CH<sub>4</sub>) from terrestrial ecosystems could increase or decrease in response to climate change, thereby either accelerating or decelerating global warming
- Field experiments examining the effects of climate change on N<sub>2</sub>O and CH<sub>4</sub> emissions provide important information that may help improve long-term predictions with process-based models
- A rise in atmospheric carbon dioxide concentration often increased N<sub>2</sub>O emissions in fertilized systems and CH<sub>4</sub> emissions in wetlands, peatlands, and rice paddy fields; such increases may enhance global warming
- Conversely, responses of N<sub>2</sub>O and CH<sub>4</sub> emissions to elevated temperatures have been inconsistent in many ecosystems

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Controlled field experiments that manipulate atmospheric CO<sub>2</sub> levels and temperatures allow for a systematic evaluation of ecosystem responses. Experimental manipulations of CO<sub>2</sub> and temperature cause secondary changes in other environmental factors – such as soil temperature and moisture, and soil carbon (C) and nitrogen (N) availability – and also affect plant growth, microbial growth, and community composition (Pendall *et al.* 2004; Engel *et al.* 2009; Morgan *et al.* 2011). Elevated CO<sub>2</sub> and temperature effects on N<sub>2</sub>O and CH<sub>4</sub> fluxes can therefore be investigated in a holistic way that incorporates all of these changes. Furthermore, manipulative field experiments allow for single, combined, and interactive effects of elevated CO<sub>2</sub> and temperature to be investigated. However, these field manipulations are usually performed with step increases in CO<sub>2</sub> concentration and temperature that may cause different effects than would gradual increases (Klironomos *et al.* 2005). Given the large monetary costs associated with maintaining such treatments in the field, these experiments usually do not extend for more than 5 years, which leaves much uncertainty concerning the long-term effects of these treatments. Despite these limitations, manipulative field experiments provide important information about the effects of elevated CO<sub>2</sub> and temperature on N<sub>2</sub>O and CH<sub>4</sub> fluxes under realistic conditions that may help improve long-term predictions with process-based models.

Here, we summarize results from manipulative field experiments conducted in different terrestrial ecosystems and assess how elevated CO<sub>2</sub> and temperature affected soil fluxes of N<sub>2</sub>O and CH<sub>4</sub>. Although precipitation has a major effect on N<sub>2</sub>O and CH<sub>4</sub> fluxes (eg Borken *et al.* 2000; Goldberg and Gebauer 2009), current projections about precipitation regimes in response to climate change remain uncertain (Meehl *et al.* 2007). Meta-analysis is often applied to summarize results from independent experiments where effect sizes of individual experiments are standardized by log response ratios or differences between treatment and control groups divided by the within-group standard deviation (Hedges *et al.* 1999). However, we focus on reporting absolute effects of elevated CO<sub>2</sub> and temperature from individual studies, which allows us to (1) assess the biogeochemical importance of elevated CO<sub>2</sub> and temperature effects on N<sub>2</sub>O and CH<sub>4</sub> fluxes, and (2) relate the variability in responses among studies to site-specific soil characteristics (eg soil texture and pH).

## ■ Methods

We reviewed 41 peer-reviewed publications that reported effects on N<sub>2</sub>O and/or CH<sub>4</sub> fluxes from elevated CO<sub>2</sub> and/or temperature treatments from 45 field sites that encompass a wide range of ecosystems (WebTable 1). Most field studies manipulating atmospheric CO<sub>2</sub> used open-top chamber (OTC) or free air carbon dioxide enrichment technology. Researchers manipulated the temperatures of field plots passively, using OTC or area

covers or actively using heating cables buried in the soil or else infrared heaters installed above the canopy. Atmospheric CO<sub>2</sub> levels in elevated treatments ranged between 470 and 700 parts per million, and “warming” treatments resulted in temperature increases of between 1° and 5°C above ambient soil, canopy, or air temperatures. These atmospheric CO<sub>2</sub> and temperature increases are consistent with Intergovernmental Panel on Climate Change projections for the middle or end of the 21st century (Meehl *et al.* 2007).

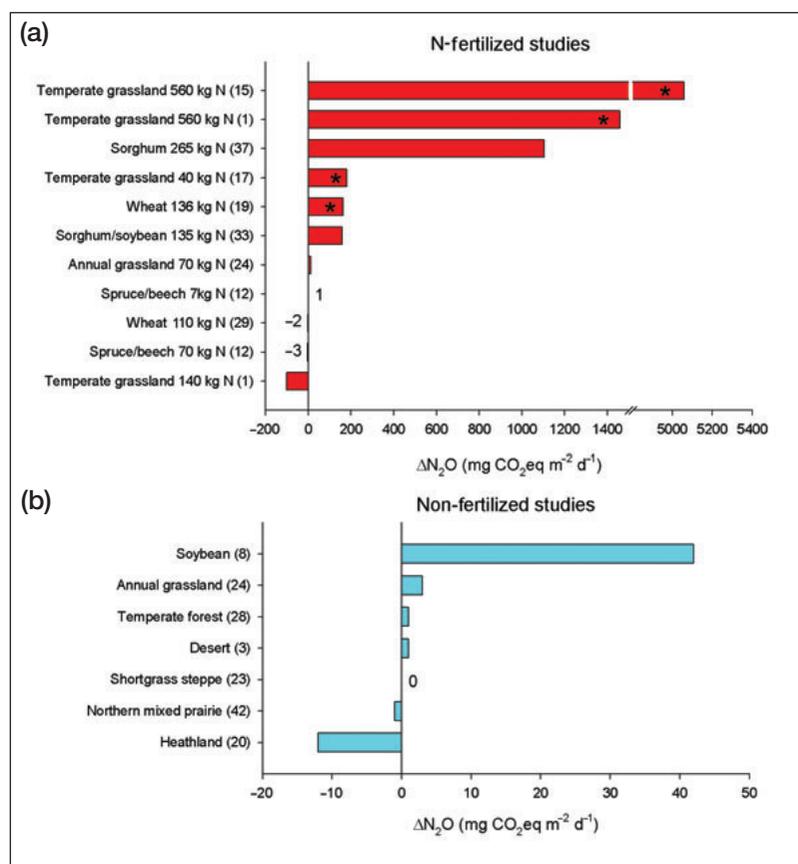
We report effects of elevated CO<sub>2</sub> and temperature on N<sub>2</sub>O and CH<sub>4</sub> fluxes as the change in the average flux rates in CO<sub>2</sub> or temperature treatments as compared with ambient control treatments during the time frame of measurement. Effects that increased or decreased emissions into the atmosphere are presented as positive or negative values, respectively. In studies where other treatments were included (eg irrigation, plant species, ozone), we averaged results across those treatments. In elevated CO<sub>2</sub> studies that included N fertilization, elevated CO<sub>2</sub> effects on N<sub>2</sub>O fluxes were reported for each level of N fertilization. All fluxes reported here are expressed in milligrams of CO<sub>2</sub> equivalents per square meter per day (mg CO<sub>2</sub>eq m<sup>-2</sup> d<sup>-1</sup>) to allow for comparison between N<sub>2</sub>O and CH<sub>4</sub> flux responses.

Because elevated CO<sub>2</sub> and temperature effects on N<sub>2</sub>O and CH<sub>4</sub> flux rates were highly variable among studies, we tested whether this variability could be explained by soil properties at each study site. We chose clay content and pH because these two factors (1) can have important effects on biological activity and N<sub>2</sub>O and CH<sub>4</sub> fluxes (Stehfest and Bouwman 2006; Fierer *et al.* 2009) and (2) are frequently reported in the literature. We related site-specific clay content and pH to site-specific changes in N<sub>2</sub>O and CH<sub>4</sub> fluxes in response to elevated CO<sub>2</sub> and temperature. At some sites, N<sub>2</sub>O and CH<sub>4</sub> fluxes were reported in more than one study during different time periods; in those cases, the fluxes were averaged across the different studies and weighted by the time period of measurement. Using JMP (version 8.0.1; SAS Institute, Cary, North Carolina), we ran linear and non-linear regressions where flux measurements conducted over longer time periods were more heavily weighted (Dijkstra and Morgan 2012).

## ■ Results

### *Effects of elevated CO<sub>2</sub> on N<sub>2</sub>O fluxes*

Elevated CO<sub>2</sub> levels had highly variable effects on N<sub>2</sub>O fluxes (Figure 1). The largest increases in N<sub>2</sub>O emissions in response to CO<sub>2</sub> treatments were observed in N-fertilized studies (up to 5058 mg CO<sub>2</sub>eq m<sup>-2</sup> d<sup>-1</sup>) and were frequently significant (Figure 1a). In contrast, the effects of elevated CO<sub>2</sub> on N<sub>2</sub>O were consistently non-significant in non-fertilized studies (Figure 1b). In a meta-analysis that included growth chamber and greenhouse studies, van Groenigen *et al.* (2011) concluded that elevated CO<sub>2</sub>



**Figure 1.** Effects of elevated CO<sub>2</sub> on N<sub>2</sub>O emissions in (a) N-fertilized and (b) non-fertilized studies. Effects of elevated CO<sub>2</sub> are expressed in mg CO<sub>2</sub>eq m<sup>-2</sup> d<sup>-1</sup> ( $\Delta N_2O$  = absolute difference in the N<sub>2</sub>O flux between elevated and ambient CO<sub>2</sub>). Bars with asterisks indicate that the effect of elevated CO<sub>2</sub> was significant ( $P < 0.05$ ). When N addition was included as a treatment, effects of elevated CO<sub>2</sub> are shown for each N-addition level, while amounts of N fertilizer (in kg N ha<sup>-1</sup> yr<sup>-1</sup>) are included on the y axis. If other treatments were included in the study, effects of elevated CO<sub>2</sub> were calculated across those treatments. Numbers in parentheses after the ecosystem type and fertilizer amount refer to reference numbers in WebTable 1.

significantly increased N<sub>2</sub>O emissions by 19%. Our results suggest that emissions-related effects of elevated CO<sub>2</sub>, in combination with N fertilization, may be intensified. Indeed, the three largest increases in N<sub>2</sub>O emissions under elevated CO<sub>2</sub> occurred in studies with the highest N-fertilization rates (ranging from 265 to 560 kg N ha<sup>-1</sup> yr<sup>-1</sup>; Figure 1a).

Nitrous oxide fluxes were often measured during the growing season, which includes high-emission periods after N-fertilizer applications as compared with periods of lower emissions during the winter (Stehfest and Bouwman 2006). If CO<sub>2</sub> effects on N<sub>2</sub>O fluxes were smaller during the winter than during the growing season, then CO<sub>2</sub> effects – when considered over the course of a given year – are lower than reported here.

Increased N<sub>2</sub>O emissions can result from elevated CO<sub>2</sub> because of the effects of the latter on soil moisture, labile C, or both (Ineson *et al.* 1998; Kammann *et al.* 2008; Niboyet *et al.* 2011). Elevated CO<sub>2</sub> often increases soil

moisture as a result of reduced plant stomatal conductance and leaf transpiration, which increases plant water-use efficiency (Morgan *et al.* 2011). Furthermore, elevated CO<sub>2</sub> often increases labile C input as a result of the increased belowground plant-C allocation (Rogers *et al.* 1994; Milchunas *et al.* 2005). Higher soil moisture levels can create anaerobic conditions that are conducive to denitrification and N<sub>2</sub>O emissions, and labile C input is an important energy source for denitrifiers (Firestone 1982). The fact that CO<sub>2</sub>-induced increases in N<sub>2</sub>O emissions only occur with N-fertilization suggests that N<sub>2</sub>O production is also often limited by inorganic N availability in terrestrial ecosystems. Indeed, without N fertilization, increased plant demand for N under elevated CO<sub>2</sub> could reduce N availability for nitrifiers and denitrifiers, thereby constraining elevated CO<sub>2</sub> effects on N<sub>2</sub>O emissions (Hungate *et al.* 1997; Mosier *et al.* 2002). Thus, elevated CO<sub>2</sub> conditions increase N<sub>2</sub>O emissions only when N fertilizer is applied in excess of plant N demand.

The wide variability in N<sub>2</sub>O emissions observed in response to elevated CO<sub>2</sub> in fertilized systems could be partially explained by site-specific differences in soil clay content. Sites differed in their clay content by between 6% and 34%, and a significant positive relationship ( $r^2 = 0.78$ ,  $P = 0.002$ ) was detected between elevated CO<sub>2</sub> effects on N<sub>2</sub>O emissions and soil clay content in N-fertilized systems (Figure 2a). This positive relationship suggests that elevated CO<sub>2</sub> increased N<sub>2</sub>O emissions more in clayey than in sandy soils. Although this relationship was derived from a small sample size ( $n = 9$ ), it is notable, given that each study site differed in many aspects besides soil texture (eg species, management type, climate, methods). We observed no relationship between soil pH and CO<sub>2</sub> effects on N<sub>2</sub>O emissions.

#### Effects of elevated CO<sub>2</sub> on CH<sub>4</sub> fluxes

Elevated CO<sub>2</sub> effects on CH<sub>4</sub> fluxes were highly variable in upland soils (Figure 3a). Soils in upland studies were predominantly net sinks for CH<sub>4</sub> (through CH<sub>4</sub> oxidation by methanotrophic bacteria). Thus, increases and decreases in CH<sub>4</sub> uptake are shown as negative and positive effects, respectively, in Figure 3a. Significantly elevated CO<sub>2</sub> effects, all of which were positive (ie decreased CH<sub>4</sub> uptake), were observed in only three studies. An increase in soil moisture under elevated CO<sub>2</sub> may have either reduced CH<sub>4</sub> diffusion into the soil (thereby reducing the amount of CH<sub>4</sub> oxidation by methanotrophs) or increased

CH<sub>4</sub> production by methanogens (Phillips *et al.* 2001).

Elevated CO<sub>2</sub> often increased CH<sub>4</sub> emissions in wetlands, peatlands, and rice paddy fields (Figure 3, b and c). Significant increases in CH<sub>4</sub> emissions in response to elevated CO<sub>2</sub> were observed in one marsh and in four rice paddy studies; these increases were much larger than those observed in uplands. The anoxic conditions in wetlands, peatlands, and rice paddies promote the production of CH<sub>4</sub> by methanogenic bacteria. Increased CH<sub>4</sub> production in these systems, when subjected to elevated CO<sub>2</sub> conditions, has been attributed to increased C input into the soil (Ziska *et al.* 1998; Tokida *et al.* 2010). As with denitrifiers producing N<sub>2</sub>O, methanogens require organic C to produce CH<sub>4</sub>, and elevated CO<sub>2</sub> may fuel methanogens with greater inputs of belowground C to increase CH<sub>4</sub> production.

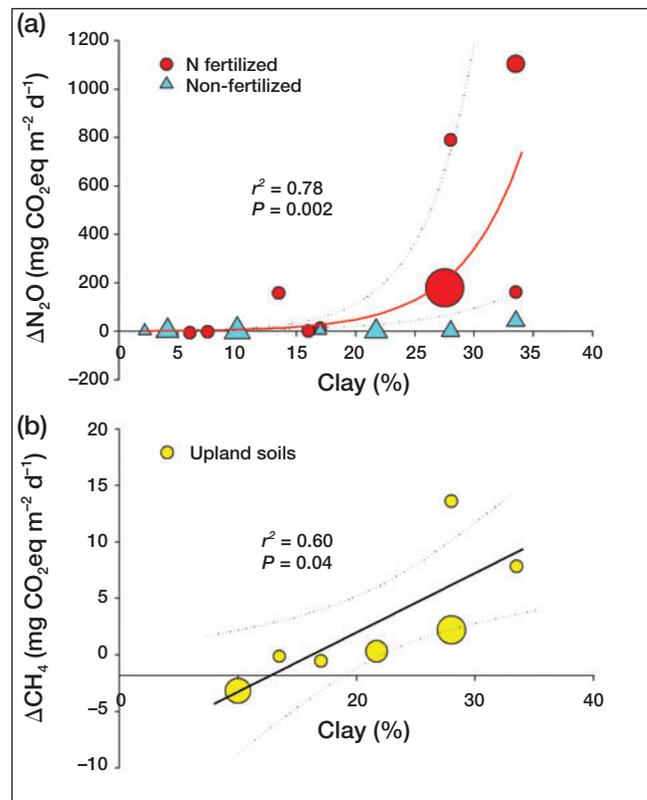
In terms of CO<sub>2</sub>eq, the effects of elevated CO<sub>2</sub> on CH<sub>4</sub> production in rice paddies were of similar magnitude to the effects of elevated CO<sub>2</sub> on N<sub>2</sub>O emissions in the high-N-fertilized systems. Considering that most rice paddies are also fertilized with N, N<sub>2</sub>O emissions in response to elevated CO<sub>2</sub> conditions may be substantial in these systems. However, none of the rice paddy studies reported any effects of elevated CO<sub>2</sub> on N<sub>2</sub>O emissions. Nevertheless, rice paddy fields appear to be one of the more sensitive ecosystems in terms of how non-CO<sub>2</sub> GHG emissions respond to elevated CO<sub>2</sub>.

Similar to the effects of elevated CO<sub>2</sub> on N<sub>2</sub>O emissions, the effect of elevated CO<sub>2</sub> on net CH<sub>4</sub> emissions in upland soils increased with clay content ( $r^2 = 0.60$ ,  $P = 0.04$ ; Figure 2b) but was not related to soil pH. As with the relationship for N<sub>2</sub>O, the associated number of data points was small ( $n = 7$ ). Greater sample sizes are therefore necessary to test the robustness of these relationships.

Soil texture largely determines the water-holding capacity and pore-size distribution in soils. Clayey soils have more micropores than sandy soils, and are therefore able to hold water more strongly; thus, anoxic conditions conducive to N<sub>2</sub>O and CH<sub>4</sub> production may be more easily created and maintained in clayey soils (Stehfest and Bouwman 2006). Any changes in soil moisture caused by elevated CO<sub>2</sub> may therefore alter N<sub>2</sub>O and CH<sub>4</sub> production to a greater degree in clayey soils than in sandy soils. Similarly, increased soil moisture may also decrease the diffusivity of CH<sub>4</sub> into soils more rapidly in clayey than in sandy soils (Thorbjørn *et al.* 2008). As such, clayey soils may be more sensitive to elevated CO<sub>2</sub> in terms of N<sub>2</sub>O and CH<sub>4</sub> production.

### Effects of increased temperature on N<sub>2</sub>O fluxes

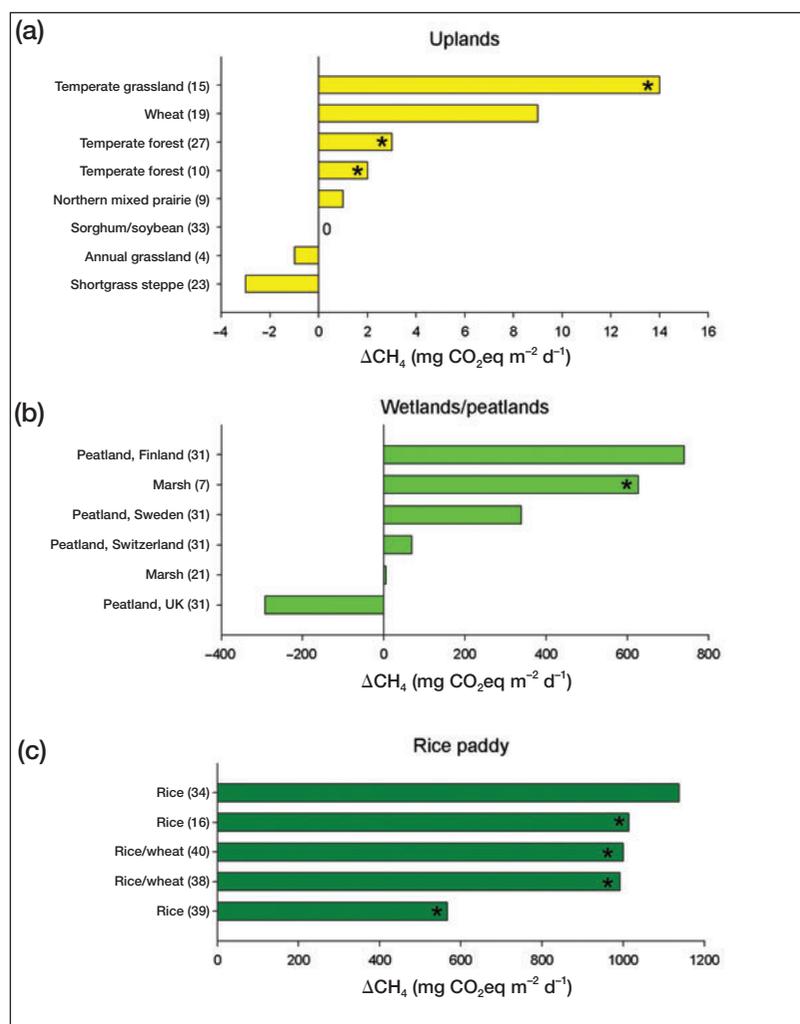
As with elevated CO<sub>2</sub>, increased temperatures affected N<sub>2</sub>O emission fluxes variably, ranging from a decrease of 111 mg CO<sub>2</sub>eq m<sup>-2</sup> d<sup>-1</sup> to an increase of 56 mg CO<sub>2</sub>eq m<sup>-2</sup> d<sup>-1</sup> (Figure 4). Significant positive and negative effects of warming were observed in both N-fertilized and non-fertilized settings. The effects of elevated temperature on



**Figure 2.** The effect of elevated atmospheric CO<sub>2</sub> concentration on (a) N<sub>2</sub>O and (b) CH<sub>4</sub> fluxes as a function of clay content. Each data point represents the result of one study or several studies at a specific site and soil clay content. For N<sub>2</sub>O, results were separated for sites with (red circles) or without (blue triangles) N fertilization. For CH<sub>4</sub>, only upland sites were included in the relationship. The size of the data points represents the weight (duration of measurement) used in the regression. Dotted lines represent 95% confidence bands.

N<sub>2</sub>O emissions were inconsistent and remained relatively small, even in the presence of N fertilization, when compared with those of elevated CO<sub>2</sub>.

There are several possible reasons for this outcome. First, elevated temperatures affect multiple processes, some of which may offset N<sub>2</sub>O emissions and result in an overall small net effect. For instance, increases in soil temperature can directly stimulate nitrifiers and denitrifiers that produce N<sub>2</sub>O, but more rapid soil drying associated with warmer conditions would have the opposite effect (McHale *et al.* 1998; Bijoor *et al.* 2008). Temperature increases could also stimulate plant growth and N uptake, thereby reducing the chance of N being lost as N<sub>2</sub>O. On the other hand, warming could boost N<sub>2</sub>O emissions as a result of increased microbial activity and N supply through increased N mineralization. Second, warming often has no effect on, or sometimes even decreases, belowground C input (Dieleman *et al.* 2012). If N<sub>2</sub>O is mainly produced by denitrifiers that are C-limited, then warming conditions would have little effect. Third, in the field experiments, soil, air, or canopy temperatures were increased by 1–5°C. Unlike elevated CO<sub>2</sub> manipulations,



**Figure 3.** Effects of elevated  $\text{CO}_2$  on net  $\text{CH}_4$  emissions in (a) uplands, (b) wetlands/peatlands, and (c) rice paddy fields. Effects of elevated  $\text{CO}_2$  are expressed in  $\text{mg CO}_2\text{eq m}^{-2} \text{d}^{-1}$  ( $\Delta\text{CH}_4$  = absolute difference in the  $\text{CH}_4$  flux between elevated and ambient  $\text{CO}_2$ ). Bars with asterisks indicate that the effect of elevated  $\text{CO}_2$  was significant ( $P < 0.05$ ). If other treatments were included in the study, effects of elevated  $\text{CO}_2$  were calculated across those treatments. Numbers in parentheses after the ecosystem type refer to reference numbers in WebTable 1.

where the  $\text{CO}_2$  concentration is often doubled, these temperature increases are relatively small for most sites where  $\text{N}_2\text{O}$  fluxes were reported and, as such, effects due to warming may also be minor. We observed no relationship between the  $\text{N}_2\text{O}$  flux in response to elevated temperatures and the soil clay content or pH at each site, possibly because of the complex effects of warming on  $\text{N}_2\text{O}$ .

#### Effects of increased temperature on $\text{CH}_4$ fluxes

Warming treatments increased net  $\text{CH}_4$  uptake (ie resulted in a more negative  $\text{CH}_4$  flux) in most upland studies and had variable effects on net  $\text{CH}_4$  emissions in wetlands, peatlands, and rice paddy fields (Figure 5). The increase in  $\text{CH}_4$  uptake observed with warming has been associated with the direct effects of higher soil temperatures on  $\text{CH}_4$  oxidation and lower soil moisture content,

which increases diffusivity of  $\text{CH}_4$  into the soil (Peterjohn *et al.* 1994; Sjögersten and Wookey 2002). In contrast with elevated  $\text{CO}_2$ , we found no relationship between the effects of elevated temperature on  $\text{CH}_4$  uptake and soil clay content. This absence of a significant relationship may be due to the much smaller range in clay content among sites that underwent warming treatments (between 15% and 22%) and those experiencing elevated  $\text{CO}_2$  treatments. The wide variability in  $\text{CH}_4$  emissions among elevated temperature studies conducted in peatlands and rice paddies could be attributed to the variable effects of warming on root biomass production and aerobic decomposition in these systems. In rice paddies subjected to warming treatments, increased  $\text{CH}_4$  emissions were associated with increased root biomass production in one study (Tokida *et al.* 2010); however, in two other studies (Ziska *et al.* 1998; Yun *et al.* 2012), root biomass production and  $\text{CH}_4$  emissions were unaffected by warming. In contrast, reduced  $\text{CH}_4$  emissions with warming treatments conducted in a peatland in Sweden were associated with faster decomposition of plant material during aerobic soil conditions (Eriksson *et al.* 2010).

In peatlands at high latitudes,  $\text{CH}_4$  emissions could further be affected by changes in the water table. As a result of climate warming, permafrost thawing could either decrease the water table (through increased drainage of melted water) or increase the water table (through thermokarst formation and flooding; Smith *et al.* 2005; Zona *et al.* 2009). For instance, a rise in the water table can promote anaerobic conditions in the soil and therefore increase  $\text{CH}_4$  production by methanogens. Indeed, in an Alaskan peatland, an increase in the water table had a greater effect on  $\text{CH}_4$  emissions than did direct warming (Turetsky *et al.* 2008).

#### Combined effects of elevated $\text{CO}_2$ and temperature

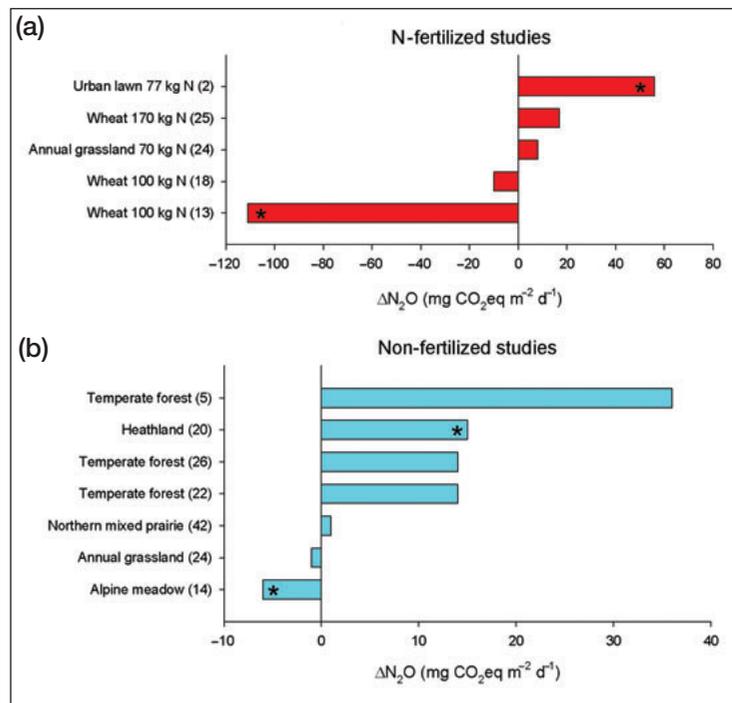
Although a considerable amount of data has been gathered regarding impacts of either elevated  $\text{CO}_2$  or elevated temperature on  $\text{N}_2\text{O}$  and  $\text{CH}_4$  emissions, little is known about their combined impacts. It is not known whether the combined effects will be equal to (additive), greater than (synergistic), or smaller than (antagonistic) the sum of single effects. Synergistic and antagonistic effects could occur when microbial processes resulting in  $\text{N}_2\text{O}$  and  $\text{CH}_4$  emissions are simultaneously controlled by more than one driver. For example, greater C inputs under ele-

vated CO<sub>2</sub> conditions and accelerated N mineralization associated with elevated temperatures could mitigate both C and N limitations for denitrification and result in synergistic effects on N<sub>2</sub>O emissions. Likewise, antagonistic effects could occur when N limitation for denitrification (under elevated CO<sub>2</sub> conditions) shifts to water limitation (under elevated temperatures), so that when both factors are combined, N<sub>2</sub>O emissions will still be constrained by one of the drivers. Combined effects could also be non-additive when one or more drivers exhibit non-linear relationships with GHG emissions (Zhou *et al.* 2008).

In the few experimental studies where interactive effects of elevated CO<sub>2</sub> and temperature were examined, no significant interactive effects on fluxes of N<sub>2</sub>O were found (Larsen *et al.* 2011; Niboyet *et al.* 2011) or CH<sub>4</sub> (Ziska *et al.* 1998; Blankinship *et al.* 2010; Tokida *et al.* 2010; Dijkstra *et al.* 2011). However, the lack of interactive effects from these experiments may be related to inadequate statistical power or to the length of time before interactive effects are expressed being longer than the duration of the experiments (Norby and Luo 2004). Modeling approaches, on the other hand, have revealed important interactive effects of climate change on heterotrophic respiration and other ecosystem processes (Luo *et al.* 2008; Zhou *et al.* 2008). Long-term field observations are needed to understand interactive effects of elevated CO<sub>2</sub> and temperature on N<sub>2</sub>O and CH<sub>4</sub> fluxes.

### ■ Critical knowledge gaps

First, there is much uncertainty regarding the effects of elevated temperatures on CH<sub>4</sub> emissions in wetlands and peatlands. For peatlands at high latitudes in particular, CH<sub>4</sub> fluxes can be sensitive to warming as a result of permafrost thawing (Schuur and Abbott 2011); this affects geomorphic and hydrological processes (McGuire *et al.* 2010) and causes large-scale spatial and temporal variations in anaerobic and aerobic soil conditions. These complex effects are almost impossible to manipulate in small-scale field experiments, although attempts have been made (Turetsky *et al.* 2008). Clearly, additional field research is needed to better understand the complex effects of elevated temperatures on CH<sub>4</sub> emissions in high-latitude soils. Second, in tropical and sub-tropical systems, there is a noted absence of field experiments, yet N cycling and N<sub>2</sub>O emissions in these systems can be extensive (Hedin *et al.* 2009); consequently, the effects of elevated CO<sub>2</sub> and temperature on N<sub>2</sub>O emissions may also be important. Third, the effects of elevated CO<sub>2</sub> and temperature on N<sub>2</sub>O fluxes in rice paddies, wetlands, or peatlands are unknown. However, N<sub>2</sub>O emissions from – and the effects of elevated CO<sub>2</sub> and temperature on – these soils (particularly those with N fertilizer additions)

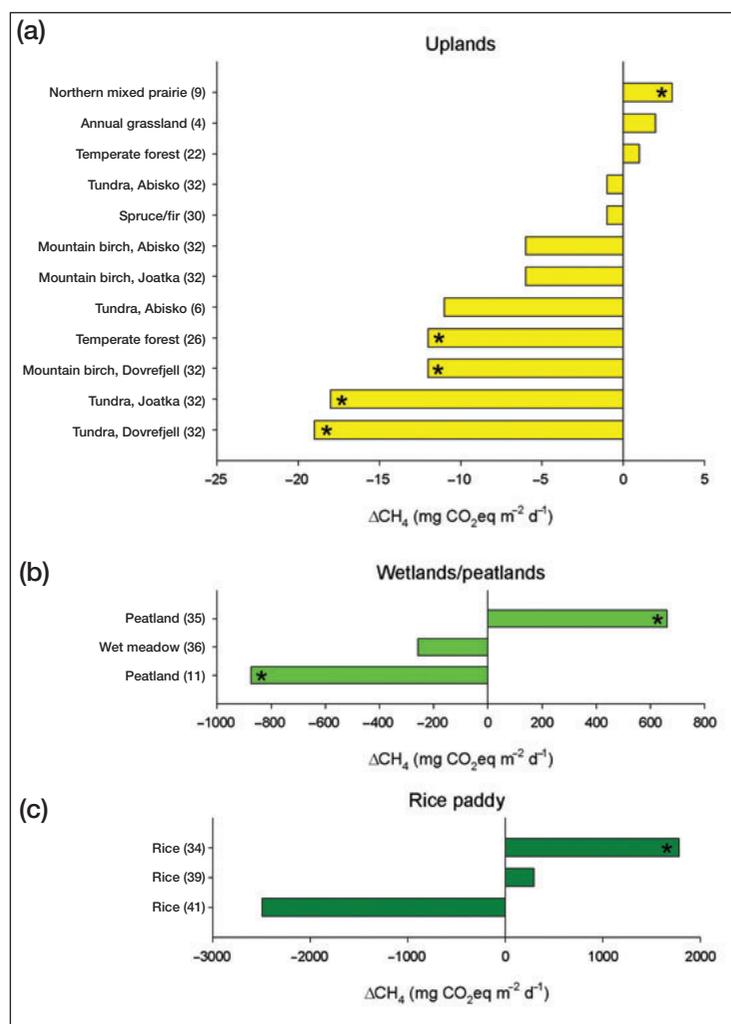


**Figure 4.** Effects of elevated temperature on N<sub>2</sub>O emissions in (a) N-fertilized and (b) non-fertilized studies. Effects of warming are expressed in mg CO<sub>2</sub>,eq m<sup>-2</sup> d<sup>-1</sup> ( $\Delta N_2O$  = absolute difference in the N<sub>2</sub>O flux between elevated and ambient temperature). Bars with asterisks indicate that the effect of elevated CO<sub>2</sub> was significant ( $P < 0.05$ ). When N addition was included, effects of warming are shown for each N-addition level, while amounts of N fertilizer (in kg N ha<sup>-1</sup> yr<sup>-1</sup>) are included on the y axis. If other treatments were included in the study, effects of warming were calculated across those treatments. Numbers in parentheses after the ecosystem type and fertilizer amount refer to reference numbers in WebTable 1.

could potentially be substantial, given that the soils are usually under anaerobic conditions. Finally, although N<sub>2</sub>O and CH<sub>4</sub> fluxes under elevated CO<sub>2</sub> and temperature conditions are affected by plant species composition or presence (Verville *et al.* 1998; Billings *et al.* 2002; Eriksson *et al.* 2010), it is unclear how N<sub>2</sub>O and CH<sub>4</sub> fluxes under these conditions will be affected by changes in plant community composition.

### ■ Conclusions

The N<sub>2</sub>O and CH<sub>4</sub> fluxes measured in different ecosystems showed various responses to elevated CO<sub>2</sub> and temperature. Nitrous oxide emissions in N-fertilized systems and CH<sub>4</sub> emissions in wetlands, peatlands, and rice paddies are particularly sensitive to, and may increase with, a rise in atmospheric CO<sub>2</sub> concentration. Our results also suggest that the effects of elevated CO<sub>2</sub> on N<sub>2</sub>O and CH<sub>4</sub> are more sensitive in clayey than in sandy upland soils. Conversely, the effects of warming on N<sub>2</sub>O and CH<sub>4</sub> fluxes were often less consistent than the effects of elevated CO<sub>2</sub>. Methane emissions, and to a lesser degree N<sub>2</sub>O emissions, showed similar sensitivity to warming as to elevated CO<sub>2</sub>, but elevated temperature caused both



**Figure 5.** Effects of elevated temperature on net  $\text{CH}_4$  emissions in (a) uplands, (b) wetlands/peatlands, and (c) rice paddy fields. Effects of warming are expressed in  $\text{mg CO}_2\text{eq m}^{-2} \text{d}^{-1}$  ( $\Delta\text{CH}_4$  = absolute difference in the  $\text{CH}_4$  flux between elevated and ambient temperature). Bars with asterisks indicate that the effect of warming was significant ( $P < 0.05$ ). If other treatments were included in the study, effects of warming were calculated across those treatments. Numbers in parentheses after the ecosystem type refer to reference numbers in WebTable 1.

strong increases and decreases in  $\text{CH}_4$  and  $\text{N}_2\text{O}$  emissions in similar ecosystems. Warming showed more consistent increases in  $\text{CH}_4$  uptake from upland soils, although these increases were small. However, because the global land area covered by upland systems is about 80 and 18 times as large as the global land area covered by rice paddies and natural wetlands, respectively (van Groenigen *et al.* 2011), increased  $\text{CH}_4$  uptake associated with elevated temperature in uplands may play an important role in offsetting the warming due to other GHGs. On the basis of results from manipulative field experiments, we conclude that  $\text{N}_2\text{O}$  and  $\text{CH}_4$  emissions from N-fertilized systems in uplands, wetlands, peatlands, and rice paddies are sensitive to a rise in atmospheric  $\text{CO}_2$  concentration, thereby serving to enhance global climate change. However, it remains uncertain whether the effects of elevated temperature on

$\text{N}_2\text{O}$  and  $\text{CH}_4$  emissions from these ecosystems will cause a negative or positive feedback.

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## References

- Bijoor NS, Czimeczik CI, Pataki DE, *et al.* 2008. Effects of temperature and fertilization on nitrogen cycling and community composition of an urban lawn. *Glob Change Biol* 14: 2119–31.
- Billings SA, Schaeffer SM, and Evans RD. 2002. Trace N gas losses and N mineralization in Mojave desert soils exposed to elevated  $\text{CO}_2$ . *Soil Biol Biochem* 34: 1777–84.
- Blankinship JC, Brown JR, Dijkstra P, *et al.* 2010. Effects of interactive global changes on methane uptake in an annual grassland. *J Geophys Res* 115: G02008.
- Borken W, Brumme R, and Xu YJ. 2000. Effects of prolonged soil drought on  $\text{CH}_4$  oxidation in a temperate spruce forest. *J Geophys Res-Atmos* 105: 7079–88.
- Dieleman W, Vicca S, Dijkstra FA, *et al.* 2012. Simple additive effects are rare: a quantitative review of plant biomass and soil process responses to combined manipulations of  $\text{CO}_2$  and temperature. *Glob Change Biol* 18: 2681–93.
- Dijkstra FA and Morgan JA. 2012. Elevated  $\text{CO}_2$  and warming effects on soil carbon sequestration and greenhouse gas exchange in agroecosystems: a review. In: Liebig MA, Follett RF, and Franzluebbers AJ (Eds). *Managing agricultural greenhouse gases*. Amsterdam, the Netherlands: Elsevier.
- Dijkstra FA, Morgan JA, Von Fischer JC, *et al.* 2011. Elevated  $\text{CO}_2$  and warming effects on  $\text{CH}_4$  uptake in a semiarid grassland below optimum soil moisture. *J Geophys Res-Bioge* 116: G01007.
- Engel EC, Weltzin JF, Norby RJ, *et al.* 2009. Responses of an old-field plant community to interacting factors of elevated  $[\text{CO}_2]$ , warming, and soil moisture. *J Plant Ecol* 2: 1–11.
- Eriksson T, Öquist MG, and Nilsson MB. 2010. Effects of decadal deposition of nitrogen and sulfur, and increased temperature, on methane emissions from a boreal peatland. *J Geophys Res-Bioge* 115: G04036.
- Fierer N, Grandy AS, Six J, *et al.* 2009. Searching for unifying principles in soil ecology. *Soil Biol Biochem* 41: 2249–56.
- Firestone MK. 1982. Biological denitrification. In: Stevenson FJ (Ed). *Nitrogen in agricultural soils*. Madison, WI: Agronomy Society of America.
- Forster P, Ramaswamy V, Artaxo P, *et al.* 2007. Changes in atmospheric constituents and in radiative forcing. In: Solomon S, Qin D, Manning M, *et al.* (Eds). *Climate change 2007: the physical science basis*. Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge, UK: Cambridge University Press.
- Goldberg SD and Gebauer G. 2009. Drought turns a central European Norway spruce forest soil from an  $\text{N}_2\text{O}$  source to a transient  $\text{N}_2\text{O}$  sink. *Glob Change Biol* 15: 850–60.
- Hedges LV, Gurevitch J, and Curtis PS. 1999. The meta-analysis of response ratios in experimental ecology. *Ecology* 80: 1150–56.
- Hedin LO, Brookshire ENJ, Menge DNL, *et al.* 2009. The nitrogen paradox in tropical forest ecosystems. *Annu Rev Ecol Evol S* 40: 613–35.
- Hungate BA, Lund CP, Pearson HL, *et al.* 1997. Elevated  $\text{CO}_2$  and nutrient addition alter soil N cycling and N trace gas fluxes

- with early season wet-up in a California annual grassland. *Biogeochemistry* **37**: 89–109.
- Ineson P, Coward PA, and Hartwig UA. 1998. Soil gas fluxes of N<sub>2</sub>O, CH<sub>4</sub> and CO<sub>2</sub> beneath *Lolium perenne* under elevated CO<sub>2</sub>: the Swiss free air carbon dioxide enrichment experiment. *Plant Soil* **198**: 89–95.
- Kammann C, Müller C, Grünhage L, *et al.* 2008. Elevated CO<sub>2</sub> stimulates N<sub>2</sub>O emissions in permanent grassland. *Soil Biol Biochem* **40**: 2194–2205.
- Klironomos JN, Allen MF, Rillig MC, *et al.* 2005. Abrupt rise in atmospheric CO<sub>2</sub> overestimates community response in a model plant–soil system. *Nature* **433**: 621–24.
- Larsen KS, Andresen LC, Beier C, *et al.* 2011. Reduced N cycling in response to elevated CO<sub>2</sub>, warming, and drought in a Danish heathland: synthesizing results of the CLIMAITE project after two years of treatments. *Glob Change Biol* **17**: 1884–99.
- Luo Y, Gerten D, Le Maire G, *et al.* 2008. Modeled interactive effects of precipitation, temperature, and [CO<sub>2</sub>] on ecosystem carbon and water dynamics in different climatic zones. *Glob Change Biol* **14**: 1986–99.
- McGuire AD, Macdonald RW, Schuur EAG, *et al.* 2010. The carbon budget of the northern cryosphere region. *Curr Opin Environ Sustain* **2**: 231–36.
- McHale PJ, Mitchell MJ, Bowles FP, *et al.* 1998. Soil warming in a northern hardwood forest: trace gas fluxes and leaf litter decomposition. *Can J Forest Res* **28**: 1365–72.
- Meehl GA, Stocker TF, Collins WD, *et al.* 2007. Global climate projections. In: Solomon S, Qin D, Manning M, *et al.* (Eds). *Climate change 2007: the physical science basis. Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge, UK: Cambridge University Press.
- Milchunas DG, Morgan JA, Mosier AR, *et al.* 2005. Root dynamics and demography in shortgrass steppe under elevated CO<sub>2</sub>, and comments on minirhizotron methodology. *Glob Change Biol* **11**: 1837–55.
- Morgan JA, Lecain DR, Pendall E, *et al.* 2011. C<sub>4</sub> grasses prosper as carbon dioxide eliminates desiccation in warmed semi-arid grassland. *Nature* **476**: 202–05.
- Mosier AR, Morgan JA, King JY, *et al.* 2002. Soil–atmosphere exchange of CH<sub>4</sub>, CO<sub>2</sub>, NO<sub>x</sub>, and N<sub>2</sub>O in the Colorado short-grass steppe under elevated CO<sub>2</sub>. *Plant Soil* **240**: 201–11.
- Niboyet A, Brown JR, Dijkstra P, *et al.* 2011. Global change could amplify fire effects on soil greenhouse gas emissions. *PLoS ONE* **6**: e20105.
- Norby RJ and Luo Y. 2004. Evaluating ecosystem responses to rising atmospheric CO<sub>2</sub> and global warming in a multi-factor world. *New Phytol* **162**: 281–93.
- Pendall E, Bridgman S, Hanson PJ, *et al.* 2004. Below-ground process responses to elevated CO<sub>2</sub> and temperature: a discussion of observations, measurement methods, and models. *New Phytol* **162**: 311–22.
- Peterjohn WT, Melillo JM, Steudler PA, *et al.* 1994. Responses of trace gas fluxes and N availability to experimentally elevated soil temperatures. *Ecol Appl* **4**: 617–25.
- Phillips RL, Whalen SC, and Schlesinger WH. 2001. Influence of atmospheric CO<sub>2</sub> enrichment on methane consumption in a temperate forest soil. *Glob Change Biol* **7**: 557–63.
- Rogers HH, Runion GB, and Krupa SV. 1994. Plant responses to atmospheric CO<sub>2</sub> enrichment with emphasis on roots and the rhizosphere. *Environ Pollut* **83**: 155–89.
- Schlesinger WH. 1997. *Biogeochemistry: an analysis of global change*. 2nd edn. San Diego, CA: Academic Press.
- Schuur EAG and Abbott B. 2011. Climate change: high risk of permafrost thaw. *Nature* **480**: 32–33.
- Sjögersten S and Wookey PA. 2002. Spatio-temporal variability and environmental controls of methane fluxes at the forest-tundra ecotone in the Fennoscandian mountains. *Glob Change Biol* **8**: 885–94.
- Smith LC, Sheng Y, MacDonald GM, *et al.* 2005. Disappearing Arctic lakes. *Science* **308**: 1429.
- Stehfest E and Bouwman L. 2006. N<sub>2</sub>O and NO emission from agricultural fields and soils under natural vegetation: summarizing available measurement data and modeling of global annual emissions. *Nutr Cycl Agroecosys* **74**: 207–28.
- Thorbjørn A, Moldrup P, Blendstrup H, *et al.* 2008. A gas diffusivity model based on air-, solid-, and water-phase resistance in variably saturated soil. *Vadose Zone J* **7**: 1230–40.
- Tian HQ, Lu CQ, Chen GS, *et al.* 2012. Contemporary and projected biogenic fluxes of methane and nitrous oxide in North American terrestrial ecosystems. *Front Ecol Environ* **10**: 528–36.
- Tokida T, Fumoto T, Cheng W, *et al.* 2010. Effects of free-air CO<sub>2</sub> enrichment (FACE) and soil warming on CH<sub>4</sub> emission from a rice paddy field: impact assessment and stoichiometric evaluation. *Biogeosciences* **7**: 2639–53.
- Turetsky MR, Treat CC, Waldrop MP, *et al.* 2008. Short-term response of methane fluxes and methanogen activity to water table and soil warming manipulations in an Alaskan peatland. *J Geophys Res-Bioge* **113**: G00A10.
- van Groenigen KJ, Osenberg CW, and Hungate BA. 2011. Increased soil emissions of potent greenhouse gases under increased atmospheric CO<sub>2</sub>. *Nature* **475**: 214–16.
- Verville JH, Hobbie SE, Chapin III FS, *et al.* 1998. Response of tundra CH<sub>4</sub> and CO<sub>2</sub> flux to manipulation of temperature and vegetation. *Biogeochemistry* **41**: 215–35.
- Xu XF, Tian HQ, Zhang C, *et al.* 2010. Attribution of spatial and temporal variations in terrestrial methane flux over North America. *Biogeosciences* **7**: 3637–55.
- Xu XF, Tian HQ, Chen GS, *et al.* 2012. Multifactor controls on terrestrial N<sub>2</sub>O flux over North America from 1979 through 2010. *Biogeosciences* **9**: 1351–66.
- Yun SI, Kang BM, Lim SS, *et al.* 2012. Further understanding CH<sub>4</sub> emissions from a flooded rice field exposed to experimental warming with elevated [CO<sub>2</sub>]. *Agr Forest Meteorol* **154–55**: 75–83.
- Zhou X, Weng E, and Luo Y. 2008. Modeling patterns of nonlinearity in ecosystem responses to temperature, CO<sub>2</sub>, and precipitation changes. *Ecol Appl* **18**: 453–66.
- Ziska LH, Moya TB, Wassmann R, *et al.* 1998. Long-term growth at elevated carbon dioxide stimulates methane emission in tropical paddy rice. *Glob Change Biol* **4**: 657–65.
- Zona D, Oechel WC, Kochendorfer J, *et al.* 2009. Methane fluxes during the initiation of a large-scale water table manipulation experiment in the Alaskan Arctic tundra. *Global Biogeochem Cy* **23**: GB2013.