

Elevated CO₂ and warming effects on CH₄ uptake in a semiarid grassland below optimum soil moisture

Feike A. Dijkstra,^{1,2} Jack A. Morgan,¹ Joseph C. von Fischer,³ and Ronald F. Follett⁴

Received 8 January 2010; revised 18 October 2010; accepted 3 November 2010; published 27 January 2011.

[1] Semiarid rangelands are a significant global sink for methane (CH₄), but this sink strength may be altered by climate change. Methane uptake is sensitive to soil moisture showing a hump-shaped relationship with a distinct optimum soil moisture level. Both CO₂ and temperature affect soil moisture, but the direction of CH₄ uptake response may depend on if the system is below or above the soil moisture optimum. Most climate change studies on CH₄ uptake have been conducted in mesic environments with soil moisture levels typically above optimum, but little is known about responses in drier systems with suboptimal soil water. We studied effects of atmospheric CO₂ (ambient versus 600 ppm), and temperature (ambient versus 1.5/3.0°C warmer day/night) on CH₄ uptake during two growing seasons in a full factorial semiarid grassland field experiment in Wyoming, United States. We observed typical hump-shaped relationships between CH₄ uptake and water filled pore space. Averaged over a range of soil moisture conditions, CH₄ uptake was not affected by elevated CO₂, but significantly decreased with warming in both seasons (25% in the first and 13% in the second season). Warming showed the strongest reduction and elevated CO₂ showed the strongest increase in CH₄ uptake when soils were below optimum moisture, indicating that these effects are particularly strong when soils are dry. Thus, directional effects of elevated CO₂ and warming on CH₄ uptake in semiarid grasslands can be opposite to their effects in mesic ecosystems because semiarid grasslands are often below optimum soil moisture for methane uptake.

Citation: Dijkstra, F. A., J. A. Morgan, J. C. von Fischer, and R. F. Follett (2011), Elevated CO₂ and warming effects on CH₄ uptake in a semiarid grassland below optimum soil moisture, *J. Geophys. Res.*, 116, G01007, doi:10.1029/2010JG001288.

1. Introduction

[2] Semiarid grasslands account for approximately 11% of the global land surface [Bailey, 1979] and have been shown to be important sinks of CH₄ on a global scale, removing between 0.5 and 5.6 Tg of CH₄ from the atmosphere each year [Mosier *et al.*, 1991]. Natural seasonal variation in soil moisture strongly modulates biological activity in semiarid grasslands [Huxman *et al.*, 2004; Potts *et al.*, 2006]. Both empirical and modeling studies suggest that changes in soil moisture caused by atmospheric CO₂ enrichment and warming also strongly affect biological activity in these grasslands [Melillo *et al.*, 1993; Morgan *et al.*, 2004; Liu *et al.*, 2009]. The CH₄ sink strength of semiarid grasslands may also be sensitive to changes in soil

moisture caused by climate change, but there is still much uncertainty about how CH₄ uptake in semiarid grasslands is affected by climate change factors such as atmospheric CO₂ enrichment and warming.

[3] The rate of CH₄ uptake is sensitive to soil moisture and typically shows a hump-shaped relationship with soil moisture [Torn and Harte, 1996; Bowden *et al.*, 1998; Del Grosso *et al.*, 2000]. At high soil moisture contents the CH₄ uptake rate is limited by diffusivity of CH₄ into the soil, while very low moisture contents limit biological activity of methanotrophs [von Fischer *et al.*, 2009] (Figure 1). Soil moisture often increases under elevated CO₂ [Niklaus *et al.*, 1998; Nelson *et al.*, 2004] because of increased plant stomatal closure and increased plant water use efficiency [Morgan *et al.*, 2004], but decreases with warming [Harte *et al.*, 1995; Dermody *et al.*, 2007]. Thus opposing effects of elevated CO₂ and warming on soil moisture may also have opposing effects on CH₄ uptake. Further, the direction of the effects of elevated CO₂ and warming on CH₄ uptake may depend on if soil moisture conditions are dry causing limitation for methanotroph activity or wet causing limitation of diffusivity (Figure 1).

[4] In several field studies atmospheric CO₂ enrichment decreased CH₄ uptake [Ineson *et al.*, 1998; Ambus and Robertson, 1999; Phillips *et al.*, 2001; McLain *et al.*, 2002;

¹Rangeland Resources Research Unit, ARS, USDA, Fort Collins, Colorado, USA.

²Now at Faculty of Agriculture, Food and Natural Resources, University of Sydney, Eveleigh, New South Wales, Australia.

³Department of Biology and Natural Resource Ecology Laboratory, Colorado State University, Fort Collins, Colorado, USA.

⁴Soil, Plant, and Nutrient Research Unit, ARS, USDA, Fort Collins, Colorado, USA.

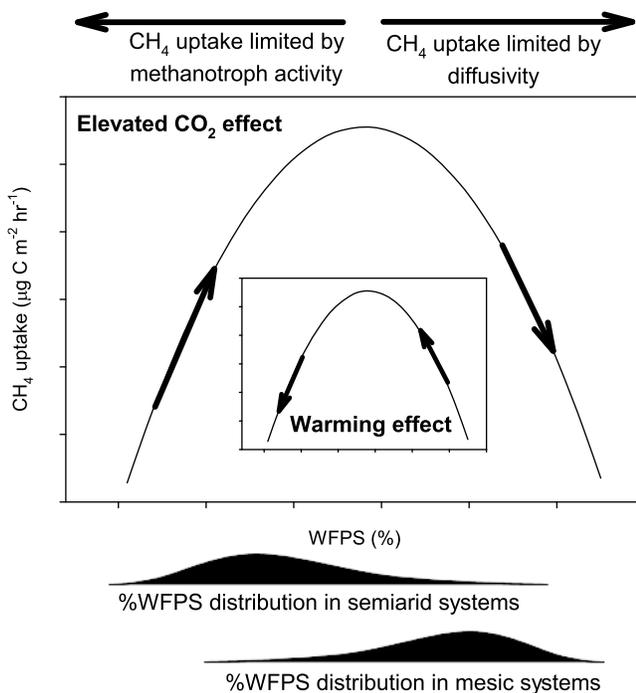


Figure 1. The relationship between CH₄ uptake and % water filled pore space (%WFPS), and how elevated CO₂ and warming affect CH₄ uptake mediated by their effects on soil moisture. Increased %WFPS under elevated CO₂ would increase CH₄ uptake when %WFPS is below optimum (i.e., when CH₄ uptake is limited by methanotroph activity) and decrease CH₄ uptake when WFPS is above optimum (i.e., when CH₄ uptake is limited by diffusivity). Inset shows opposite warming effects on CH₄ uptake mediated by its effect on soil moisture. Hypothetical distributions of %WFPS occurrences in semiarid and mesic systems are shown below the graph.

McLain and Ahmann, 2008], but no change in CH₄ uptake has also been observed [Kang et al., 2001; Mosier et al., 2002; Kettunen et al., 2005]. A decrease in CH₄ uptake in response to elevated CO₂ has been related to an increase in soil moisture reducing diffusivity [Ineson et al., 1998; Ambus and Robertson, 1999; McLain et al., 2002], but also to increased production of CH₄ by methanogens or to changes in the methanotroph/methanogen community in the soil [Phillips et al., 2001; McLain and Ahmann, 2008]. Experimental warming increased CH₄ uptake in some field studies, possibly due to a reduction in soil moisture increasing diffusivity [Peterjohn et al., 1994; Sjögersten and Wookey, 2002], but not in others [Christensen et al., 1997; McHale et al., 1998; Rustad and Fernandez, 1998]. An increase in soil temperature could also directly stimulate methanotroph activity [Castro et al., 1995; Bowden et al., 1998]. Often, temporal variation in field measurements of CH₄ uptake correlates positively with temporal fluctuations in soil temperature [van den Pol-van Dasselaar et al., 1998; West et al., 1999; Phillips et al., 2001], but it is not always clear how much this correlation is caused by direct or indirect temperature effects.

[5] Soil moisture contents in semiarid systems are often below optimum soil moisture [Mosier et al., 2008], and as a result, semiarid systems may respond very differently to

elevated CO₂ and warming than mesic sites where soil moisture contents may often be above this optimum (Figure 1). Most field studies that examined the effects on elevated CO₂ or warming on CH₄ fluxes were done in temperate/boreal forests, and (sub)arctic ecosystems where soil moisture contents may have been at or above optimum soil moisture, and thus where CH₄ uptake may have been limited by CH₄ diffusivity [Peterjohn et al., 1994; Christensen et al., 1997; McHale et al., 1998; Rustad and Fernandez, 1998; Ambus and Robertson, 1999; Phillips et al., 2001; McLain et al., 2002; Sjögersten and Wookey, 2002]. One exception is a study by Mosier et al. [2002] who studied the effects of elevated CO₂ using open top chambers on CH₄ uptake in a semiarid grassland in Colorado, United States. Five years of elevated CO₂ did not significantly alter CH₄ uptake in this system. However, CH₄ uptake tended to be greater under elevated CO₂ than under ambient CO₂, suggesting that greater soil water savings under elevated CO₂ reduced soil moisture limitation on methanotroph activity.

[6] To better understand the response of dry grassland CH₄ uptake to climate change factors, we examined the effects of elevated CO₂ and warming on CH₄ uptake during the growing season in a semiarid grassland in Wyoming, United States. We hypothesized that elevated CO₂ and warming effects on CH₄ uptake in this semiarid grassland are mediated by their effects on soil moisture. Because CH₄ uptake in this semiarid grassland is often below optimum soil moisture thereby limiting methanotroph activity, we expected that increased soil moisture under elevated CO₂ would increase CH₄ uptake, and that decreased soil moisture with warming would decrease CH₄ uptake during these dry periods.

2. Materials and Methods

2.1. Site Description and Experimental Design

[7] We did our study in the Prairie Heating And CO₂ Enrichment (PHACE) experiment located at the USDA-ARS High Plains Grasslands Research Station, Wyoming, United States (41°11' latitude, 104°54' longitude). Mean annual precipitation is 384 mm and mean air temperatures are 17.5°C in July and -2.5°C in January. The vegetation is of a northern mixed-grass prairie dominated by *Pascopyrum smithii*, *Hesperostipa comata* (C₃ grasses) and *Bouteloua gracilis* (C₄ grass). These 3 species comprise approximately 80% of the total aboveground biomass. Other species include *Carex eleocharis* (sedge), *Artemisia frigida* (subshrub), and *Sphaeralcea coccinea* (forb). The site has not been grazed since 2004. Soils are of the Ascalon (north side) and Altvan series (south side, fine-loamy, mixed, mesic Aridic Argiustoll). The well-drained soils have a pH of 7.0 (top 20 cm).

[8] Site preparation of the PHACE experiment started in 2005. Twenty-five circular plots (12 on the north side, 18 on the south side) were established by installing a 60 cm deep, 3.7 m diameter plastic perimeter barrier around each experimental plot. A steel flange buried to 25 cm into the soil divided each plot in half. One half of the plot (randomly assigned to the north or south part of the plot) was disturbed and planted with invasive weeds, while in the other half native vegetation was maintained. Our study was done on the side with the native vegetation. The CO₂ and warming treatments were established in 20 plots ("core plots") in a full

factorial design (2 levels of CO₂ * 2 levels of warming * 5 replicates). Ten plots received an elevated atmospheric CO₂ concentration of 600 ppm using free-air CO₂ enrichment technology [Miglietta *et al.*, 2001]. The CO₂ was injected into the plot from a plastic pipe, perforated with 300 μ m laser-drilled holes, surrounding the plot (diameter 3.4 m). Plots were treated with CO₂ only during the day and during the growing season (April–November), and started in April 2006. The canopy of 5 ambient CO₂ and 5 elevated CO₂ plots were warmed 1.5°C above ambient temperature during the day and 3°C above ambient temperature during the night with 1000 W ceramic infrared heaters. Each plot was heated by six heaters installed on a triangular frame 1.5 m above the ground. Heaters were controlled by a proportional-integral-derivative feedback loop [Kimball *et al.*, 2008]. The heating treatment was year-round and started in April 2007. The 5 plots not used for the CO₂ and warming were irrigated with 20 mm four times during the growing season of 2007 (total of 80 mm yr⁻¹) and three times during the growing season of 2008 (total of 60 mm yr⁻¹). We included this treatment to better understand the relationship between CH₄ uptake and soil water. Because 2006 was a dry year, all 25 plots received 20 mm irrigations eight times during the course of the season (total of 160 mm) to facilitate establishment in the adjacent invasive species experiment.

2.2. Measurements

[9] In 2005 EnviroSMART soil moisture probes were installed to 80 cm soil depth, one probe at each plot. Volumetric soil moisture was monitored at 10, 20, 40, 60, and 80 cm soil depth. In 2005 thermocouples were installed at 3 and 10 cm soil depth in each plot. Soil moisture and temperature data were logged every hour, starting in July 2006. Water filled pore space (WFPS) in the top 15 cm of the soil was calculated based on soil moisture measured at 10 cm soil depth and bulk densities measured at 0–5 and 5–15 cm soil depth in 2005.

[10] In March 2007 we pounded polyvinyl chloride (PVC) circular chamber bases (height, 10 cm; diameter, 20 cm) 8 cm into the ground, one base in each plot. From April to October in 2007 and 2008 we measured CH₄ fluxes approximately once every 2 weeks (total of 16 measurements in 2007 and 14 measurements in 2008) using a vented closed chamber technique [Hutchinson and Mosier, 1981]. Leakage of CH₄ in or out of this system is considered to be small, and thus CH₄ fluxes measured are representative of the area covered by the base. Midmornings of each sampling day a vented closed PVC chamber (height, 10 cm; diameter, 20 cm) was placed on the base in each plot, sealed off with a rubber band, and 30 ml gas samples were taken from the headspace at 0, 15, 30, and 45 min after chamber placement. Gas samples were analyzed for CH₄ on a gas chromatograph equipped with a flame ionization detector. Methane fluxes were calculated using linear regressions using the CH₄ concentrations measured in the four samples taken at 15 min intervals. We calculated cumulative CH₄ uptake during the growing season of 2007 and 2008 by multiplying the average CH₄ uptake rate between two measuring dates by the time interval between two measuring dates, and by adding the preceding CH₄ uptake. We did not measure CH₄ uptake before the treatments started to test for preexisting differences among treatments. However, because the CO₂, warming, and irrigation treat-

ments were replicated 5 times randomly assigned at our site (2 replicates on the north side and 3 on the south side), we are confident that we strongly reduced potential treatment effects caused by spatial variability of the study site.

2.3. Statistical Analyses

[11] We used repeated measures analysis of variance (repeated measures ANOVA) to test for main effects of CO₂ (ambient versus elevated), warming (no warming versus warming, both between-subjects factors), date (within-subjects factor), and their interactions on WFPS in the top 15 cm of the soil (on weekly averages) and on CH₄ uptake rates from April to October in 2007 and 2008. For the same variables we used a separate repeated measures ANOVA to test for irrigation effects (5 irrigated plots versus 5 plots under ambient CO₂ and no warming, between-subjects factor), date, and their interaction. We used ANOVA to test for main effects of CO₂, warming, and their interaction (20 core plots only), and for irrigation effects separately, on cumulative CH₄ uptake during the growing season of 2007 and 2008. We used quadratic regression analyses to examine relationships between average daily CH₄ uptake rates and WFPS in the top 15 cm of the soil and soil temperature at 3 and 10 cm soil depth. We further tested to what extent CO₂, warming, and irrigation effects on CH₄ uptake measured on each date in both years could be explained by their effects on soil moisture. We first averaged CH₄ uptake for each date (average of the 5 replicates for each treatment). We then used the average CH₄ uptake for each date as replicates in analyses of covariance (ANCOVAs) thereby removing the date effect, since we were no longer interested in *when* treatment effects occurred, but in how much of the treatment effects on CH₄ uptake could be explained by treatment-induced changes in soil moisture with WFPS as the covariate. Because CH₄ uptake showed a hump-shaped relationship with WFPS, we included a quadratic term of the covariate in the ANCOVAs. We also compared CO₂ and warming treatment effects in the ANCOVAs with their effects in ANOVAs without WFPS as the covariate. We included the random effect of soil type (north versus south, block effect) in all ANOVAs and ANCOVAs. Note that we tested the irrigation effect separately with ANOVAs and ANCOVAs using the 5 ambient CO₂ and ambient temperature plots without irrigation and the 5 ambient CO₂ and ambient temperature plots with irrigation. We log-transformed data when necessary to reduce heteroscedasticity. All statistical analyses were done with JMP (version 4.0.4).

3. Results

[12] Growing season precipitation in 2007 was close to average (from 1 April to 31 October, 315 mm fell compared to 310 mm on average from 1951 to 2008), while in 2008 it was slightly above average (354 mm). From 1 April to 31 October 2007 and 2008, WFPS to 15 cm soil depth was significantly higher under elevated than under ambient CO₂ (P = 0.003 in 2007 and in 2008, repeated measures ANOVA, Figures 2a and 2b). Warming significantly reduced WFPS in both years (P = 0.02 in 2007 and P = 0.005 in 2008). We observed no significant CO₂*warming interactions on WFPS in either year (P > 0.1). Irrigation events caused spikes in

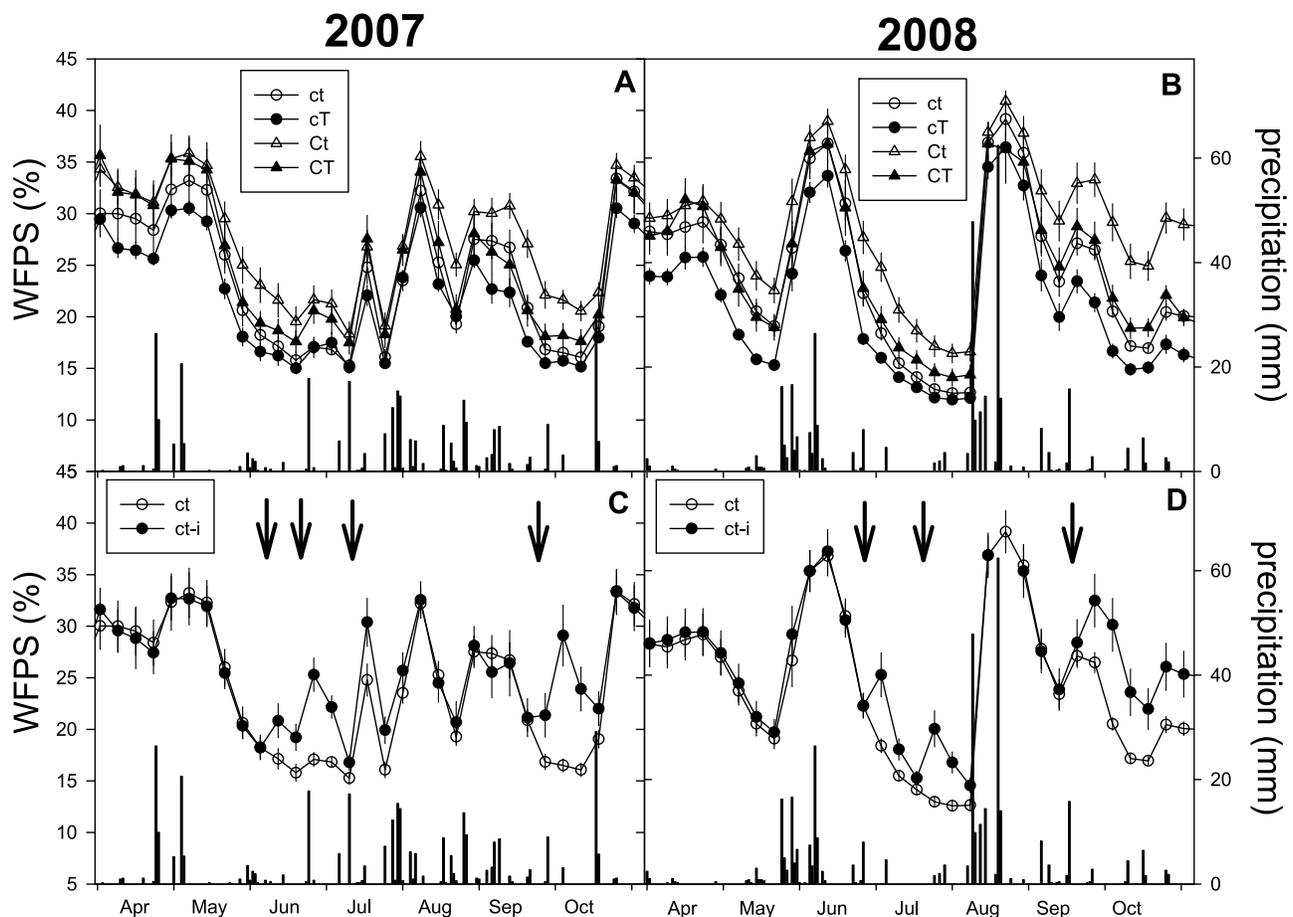


Figure 2. Water filled pore space (WFPS, left-hand axis) to 15 cm soil depth in 2007 and 2008 averaged by (a and b) elevated CO₂ treatment (ct, ambient CO₂ and ambient temperature; cT, ambient CO₂ and elevated temperature; Ct, elevated CO₂ and ambient temperature; CT, elevated CO₂ and elevated temperature) and by (c and d) irrigation treatment (ct-i, ambient CO₂ and ambient temperature, but irrigated). Each data point is a weekly average of the hourly logged data. Bars in each panel show daily precipitation (right-hand axis). Arrows in Figures 2c and 2d indicate the time when 20 mm water events occurred. Error bars indicate 1 SE.

WFPS relative to the ambient plots, but otherwise were similar in WFPS (Figures 2c and 2d).

[13] In this ecosystem methane was a net sink throughout the growing season of 2007 and 2008. We did not observe significant CO₂*warming or CO₂*warming*date interaction effects on CH₄ uptake rates in the repeated measures ANOVA. We therefore present CO₂ effects averaged across the warming treatment and warming effects averaged across the CO₂ treatment. Elevated CO₂ did not significantly affect CH₄ uptake rates in 2007 or in 2008 ($P > 0.1$, Figures 3a and 3b). However, CO₂*date interaction effects were highly significant in both years ($P < 0.0001$) with CH₄ uptake rates sometimes higher (midsummer of 2007 and 2008) and sometimes lower (late summer 2008) under elevated than under ambient CO₂. Possibly, positive CO₂ effects during midsummer when soils were relatively moist, may have canceled out negative CO₂ effects later in the season when soils were much drier. Warming significantly reduced CH₄ uptake rates in both years ($P < 0.05$, Figures 3c and 3d). This reduction occurred throughout most of the growing season, but was particularly large during the middle of the

growing season causing significant warming*date interactions in both years ($P < 0.01$). The irrigation treatment had only a marginally significant effect on CH₄ uptake rates in 2008 ($P = 0.06$, Figures 3e and 3f). The CH₄ uptake rates increased after irrigation events in midsummer, but decreased after irrigation events in late summer. Thus, not surprisingly, the irrigation*date interactions were significant in both years ($P < 0.01$).

[14] Similar to CH₄ uptake rates, elevated CO₂ had no effect on the cumulative amount of CH₄ taken up during the growing season (Table 1). Apparently, opposing effects of CO₂ on CH₄ uptake rates during different times of the growing season canceled each other out. On the other hand, warming significantly reduced the cumulative amount of CH₄ uptake by 25% in 2007 and by 13% in 2008. The cumulative amount of CH₄ uptake increased by 20% in 2007 and decreased by 20% in 2008 in response to irrigation, although the irrigation effect in 2007 was not significant.

[15] Although we found the usual hump-shaped relationship, the shape and optimum soil moisture level was affected by the CO₂ and warming treatments, suggesting that treat-

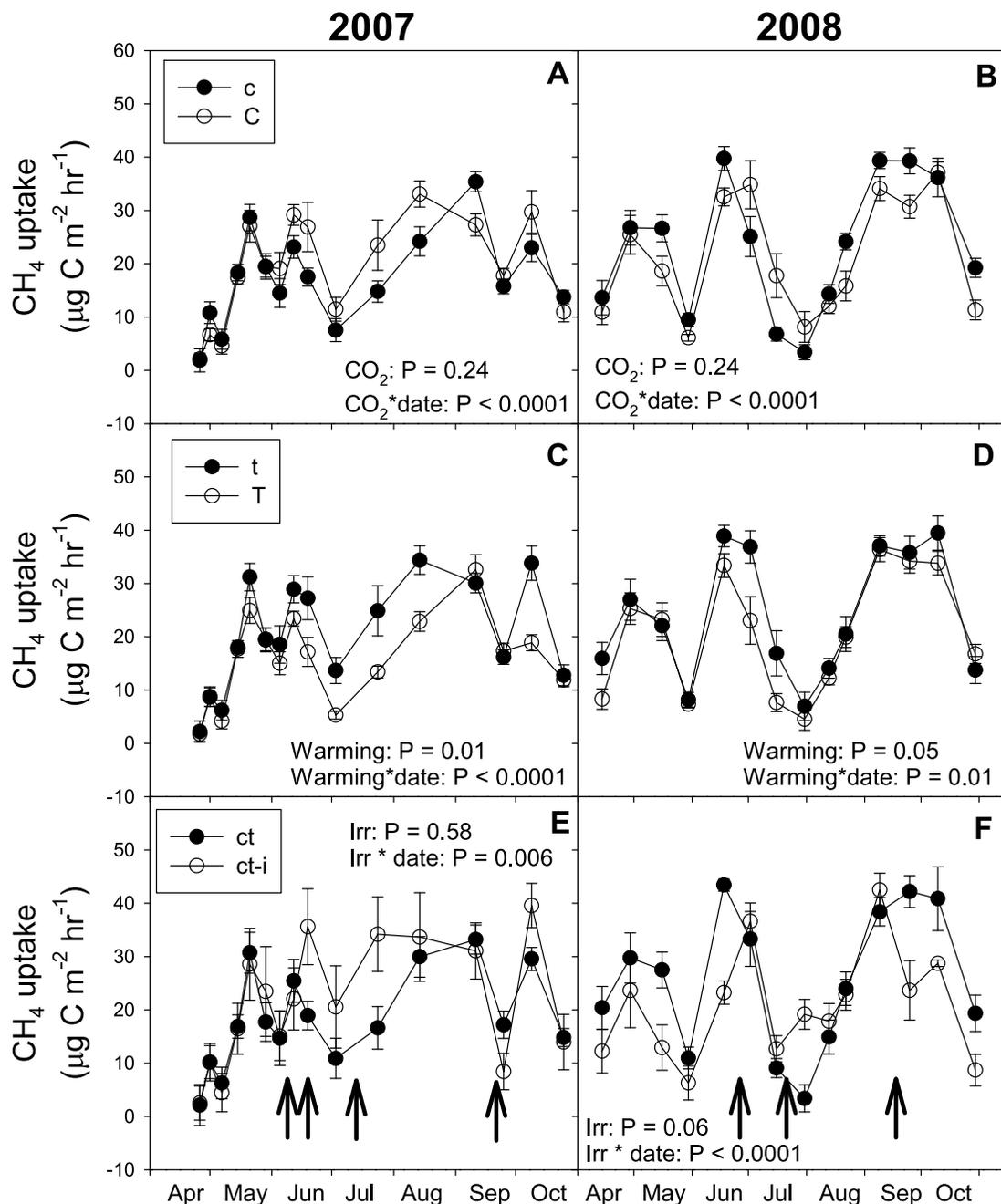


Figure 3. Methane uptake rates during the growing season of 2007 and 2008 averaged by (a and b) CO₂ treatment (c, ambient CO₂; C, elevated CO₂), by (c and d) warming treatment (t, ambient temperature; T, elevated temperature), and by (e and f) irrigation treatment (ct, ambient CO₂ and ambient temperature without irrigation; ct-i, ambient CO₂ and ambient temperature with irrigation). Arrows in Figures 3e and 3f indicate the time when 20 mm water events occurred. Error bars indicate 1 SE.

ment effects cannot solely be explained by their effects on WFPS. The CH₄ uptake rates measured in 2007 and 2008 showed a significant hump-shaped relationship with WFPS with an optimum %WFPS around 24%. This relationship with WFPS moved to the right under elevated CO₂ (averaged across warming treatment, Figure 4a) and moved downward with warming (averaged across the CO₂ treatment, Figure 4b). When we used WFPS as a covariate in the ANCOVA, then the warming treatment effect remained sig-

nificant, while there was a significant CO₂*WFPS² interaction (Table 2). This suggests that other effects than changes in WFPS caused by the CO₂ and warming treatments were involved in altering CH₄ uptake. On the other hand, the hump-shaped relationship between CH₄ uptake and WFPS was very similar between the irrigated and nonirrigated plots (Figure 4c), and there were no significant interactions with WFPS in the ANCOVA (Table 2). Soil temperature at 10 cm soil depth showed a similar hump-shaped relationship with

Table 1. Average Cumulative Amount (\pm SE) of CH₄ Uptake During the Growing Season of 2007 and 2008^a

Treatment	2007 (181 Days) (mg C m ⁻²)	2008 (199 Days) (mg C m ⁻²)
Ct	92 \pm 9	127 \pm 4
cT	75 \pm 5	105 \pm 6
Ct	113 \pm 12	112 \pm 12
CT	79 \pm 9	102 \pm 10
ct-i	111 \pm 22	103 \pm 10
ANOVA P values ^b		
CO ₂	0.13	0.23
Warming	0.006	0.05
CO ₂ *warming	0.28	0.39
Irrigation	0.44	0.04

^aHere ct, ambient CO₂ and ambient temperature; cT, ambient CO₂ and elevated temperature; Ct, elevated CO₂ and ambient temperature; CT, elevated CO₂ and elevated temperature; ct-i, ambient CO₂ and ambient temperature, but irrigated.

^bSoil type effects were never significant ($P > 0.1$) and are not reported.

CH₄ uptake, while at 3 cm soil depth this relationship was not significant (Figure 5). Soil temperature was also significantly correlated to soil moisture ($r = 0.51$ and 0.48 for relationships with soil temperature at 3 and 10 cm soil depth, respectively, $P < 0.0001$ for both relationships).

[16] The CH₄ uptake responses to elevated CO₂ and warming depended on overall dryness or wetness of the soil. We plotted the average CH₄ uptake response to elevated CO₂ (averaged across the warming treatment), and the average CH₄ uptake response to warming (averaged across the CO₂ treatment) for each measuring date in 2007 and 2008 against the average %WFPS of all 20 core plots for each date (Figure 6). Methane uptake rates were mostly higher under elevated CO₂ than ambient CO₂ when soils were below optimum %WFPS (around 24%), and mostly lower under elevated CO₂ than under ambient CO₂ when soils were above optimum %WFPS (Figure 6a). The lower CH₄ uptake rates with warming particularly occurred when soils were below optimum %WFPS (Figure 6b). Exponential curves fitted the data better than linear relationships, suggesting that both CO₂ and warming treatment effects were more sensitive under drier soil conditions. Nonirrigated soils in this semiarid system were more frequently below than above optimum %WFPS (59% of the times measured below 24% WFPS, Figure 4d).

4. Discussion

[17] Throughout the growing seasons of 2007 and 2008, CH₄ was taken up by this semiarid grassland. Methane uptake rates ranged between 2 and 44 $\mu\text{g C m}^{-2} \text{h}^{-1}$, similar to rates measured in a semiarid grassland in Colorado, United States [Mosier *et al.*, 2002, 2008] and other ecosystems [Christensen *et al.*, 1997; Ineson *et al.*, 1998; Phillips *et al.*, 2001; Sjögersten and Wookey, 2002]. As expected, the relationship between CH₄ uptake and WFPS was hump-shaped with an optimum WFPS around 24%. This optimum %WFPS agrees well with optimum %WFPS values of fine-textured sites in semiarid grasslands of Colorado [Mosier *et al.*, 2008]. Often %WFPS was below the optimum, suggesting that soil water constraints on methanotroph activity are important during the growing season.

[18] Our results indicate that CH₄ uptake in a semiarid grassland responds differently to elevated CO₂ and warming than CH₄ uptake in mesic environments. While other field studies in mesic environments have shown a decrease in CH₄ uptake in response to elevated CO₂ [Ineson *et al.*,

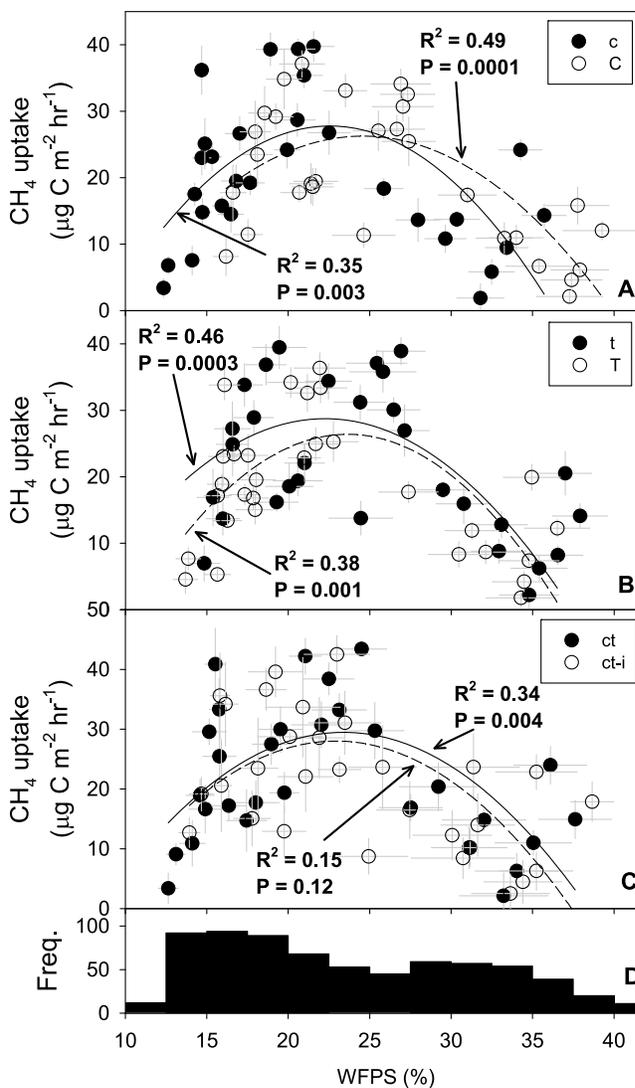


Figure 4. Methane uptake rate as a function of water filled pore space (WFPS) averaged by (a) CO₂ treatment (c and solid line, ambient CO₂; C and dashed line, elevated CO₂), by (b) warming treatment (t and solid line, ambient temperature; T and dashed line, elevated temperature), and by (c) irrigation treatment (ct and solid line, ambient CO₂ and ambient temperature without irrigation; ct-i and dashed line, ambient CO₂ and ambient temperature with irrigation), and (d) frequency distribution of %WFPS: at the time of CH₄ flux measurements in all nonirrigated plots. Each data point in Figures 4a–4c is the average CH₄ uptake and %WFPS in the ambient and elevated CO₂ plots (averaged across the warming treatment, Figure 4a), in the nonwarmed and warmed plots (averaged across the CO₂ treatment, Figure 4b), and in the ct and ct-i plots (Figure 4c) at each measuring date in 2007 and 2008. Regression R^2 and P values are based on these averaged values. Error bars are 1 SE.

Table 2. ANOVA and ANCOVA P Values for Methane Uptake During the Growing Seasons of 2007 and 2008^a

Model Effect ^b	ANOVA	ANCOVA
CO ₂	0.90	0.63
Warming	0.0001	<0.0001
WFPS		0.0003
WFPS ²		<0.0001
CO ₂ *WFPS		0.20
CO ₂ *WFPS ²		0.02
Warming*WFPS		0.64
Warming*WFPS ²		0.28
Irrigation	0.73	0.85
WFPS		0.16
WFPS ²		0.0003
Irrigation*WFPS		0.17
Irrigation*WFPS ²		0.93

^aP values are in bold when $P < 0.05$.

^bSoil type effects were never significant ($P > 0.1$) and are not reported.

1998; Ambus and Robertson, 1999; Phillips et al., 2001; McLain et al., 2002; McLain and Ahmann, 2008], we found that on average CH₄ uptake was not affected by elevated CO₂. However, CH₄ uptake did decrease under elevated CO₂ when soils were above optimum %WFPS (i.e., when CH₄ uptake was limited by diffusivity, Figure 6a). In contrast, when soils were below optimum %WFPS (i.e., when CH₄ uptake was limited by methanotroph activity), then CH₄ uptake rates tended to be higher under elevated CO₂ than under ambient CO₂. Thus, the nonsignificant main effect of CO₂ on CH₄ uptake may have been a result of negative effects during times when soils were wet cancelling out positive effects during times when soils were dry. Warming significantly reduced CH₄ uptake in both years. These results also contrast other field studies where no or positive effects of warming on CH₄ uptake have been observed [Peterjohn et al., 1994; Christensen et al., 1997; McHale et al., 1998; Rustad and Fernandez, 1998; Sjögersten and Wookey, 2002]. The reduction in CH₄ uptake with warming particularly occurred when soils were below the optimum %WFPS for CH₄ uptake

(Figure 6b), conditions that frequently happen in semiarid grasslands, but that may not happen as frequently in mesic environments. These results clearly indicate that under dry soil conditions, which frequently occur in semiarid ecosystems, elevated CO₂ and warming effects on CH₄ uptake are very different from wet soil conditions as in mesic ecosystems. The exponential relationships in Figure 6 with steeper slopes at lower %WFPS further indicate that the CO₂ and warming treatment effects become more sensitive when soil conditions become drier.

[19] Elevated CO₂ and warming effects on CH₄ uptake were largely a result of their effects on soil moisture. A higher soil moisture or %WFPS under elevated CO₂ most likely increased CH₄ uptake when in general the soils were dry (i.e., when CH₄ uptake was limited by methanotroph activity), and decreased CH₄ uptake when in general the soils were wet (i.e., when CH₄ uptake was limited by diffusivity) due to the hump-shaped relationship between CH₄ uptake rate and WFPS (Figure 1). Likewise, a lower %WFPS in the warmed plots most likely decreased CH₄ uptake rates under generally dry soil conditions suggesting a positive feedback between methane flux and climate warming [Torn and Harte, 1996]. The importance of soil moisture for CH₄ uptake responses to CO₂ and warming is further illustrated by the irrigation effects on CH₄ uptake. In 2008, irrigation significantly reduced CH₄ uptake, which was largely driven by a reduction in CH₄ uptake after the irrigation event late in the season (September) and possibly from irrigation events in 2007 carried over into the spring of 2008 when soils were relatively wet (Figure 3f, i.e., when CH₄ uptake was limited by soil moisture effects on diffusivity). Likewise, the increase in CH₄ uptake in response to irrigation events in midsummer of 2007 and 2008 (Figures 3e and 3f) occurred when soils were dry, i.e., when CH₄ uptake was limited by soil moisture effects on methanotroph activity. Thus, irrigation effects on CH₄ uptake may to a large degree depend on the timing when irrigation events occur. Similarly, expected changes in precipitation patterns for this area (less winter snow and rainfall in the winter creating drier springs) [Christensen et al., 2007;

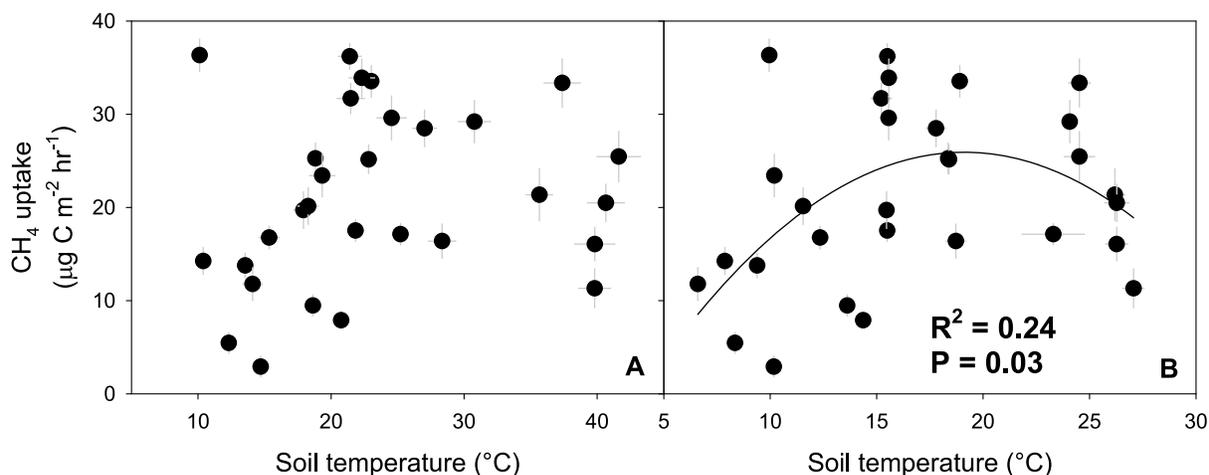


Figure 5. Methane uptake rate as a function of soil temperature at (a) 3 cm and (b) 10 cm soil depth. Each data point in each panel is the average CH₄ uptake and %WFPS of all 25 plots at each measuring date in 2007 and 2008. Regression R^2 and P values are based on these averaged values. Error bars are 1 SE.

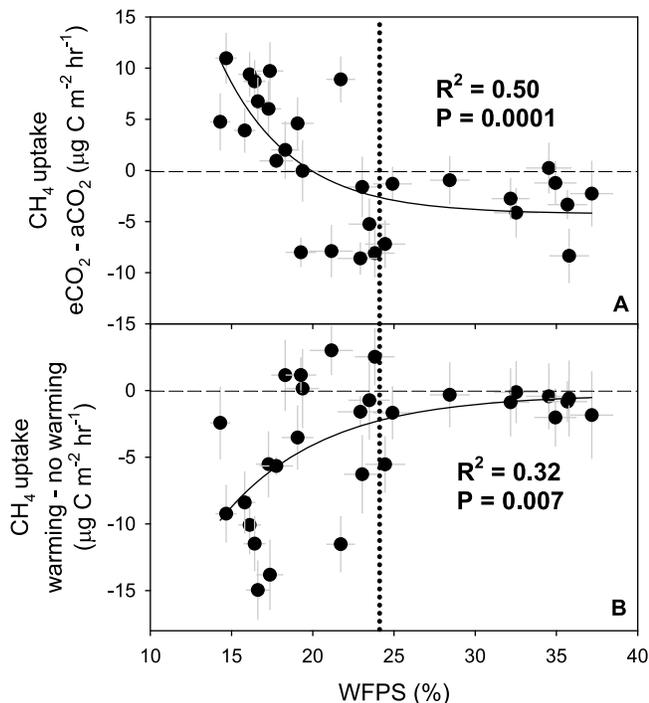


Figure 6. Methane uptake rate responses to (a) elevated CO₂ and (b) warming as a function of water filled pore space (WFPS). Each data point is the average CH₄ uptake response to elevated CO₂ (averaged across the warming treatment), and the average CH₄ uptake response to warming (averaged across the CO₂ treatment) for each measuring date in 2007 and 2008. The WFPS was calculated as the average %WFPS of all 20 core plots for each date. Regression R² and P values are based on these averaged values. Error bars are 1 SE. The vertical dotted line represents optimum %WFPS.

Overpeck and Udall, 2010] could potentially have large effects on CH₄ uptake both in magnitude and direction in these grasslands.

[20] We have little evidence that CH₄ uptake was constrained by soil temperature. When we related CH₄ uptake to soil temperature, we observed no relationship for the 3 cm soil temperatures, and a hump-shaped relationship for the 10 cm soil temperatures that was similar to, although much weaker than, the relationship between CH₄ uptake and soil moisture. Methane uptake decreased above 30°C at 3 cm soil depth and above 20°C at 10 cm soil depth. On the other hand, in a different study, the relationship between CH₄ uptake and soil temperature remained positively linear up to 35°C when soil moisture was not limiting [*Del Grosso et al.*, 2000]. It is likely that the curve-linear relationship between the 10 cm soil temperature and CH₄ uptake, and decrease in CH₄ uptake at high soil temperatures in our study was to a large extent caused by water limitation on methanotroph activity. There was a high degree of covariance between soil temperature and soil moisture, which made it difficult to separate temperature from moisture effects. Unfortunately, our study did not have a treatment with both warming and irrigation, which would have made it easier to separate soil moisture from soil temperature effects on CH₄ uptake. We note however, that the warming treatment caused a reduc-

tion in CH₄ uptake despite a significant increase in soil temperature [*Dijkstra et al.*, 2010], suggesting that, if the relationship between CH₄ uptake and soil moisture is positive, CH₄ uptake was not constrained by soil temperature in the warmed plots.

[21] We found limited support for other factors that may have contributed to the CO₂ and warming treatment effects on CH₄ uptake. The hump-shaped relationship between CH₄ uptake and WFPS under elevated CO₂ shifted slightly to the right (causing a significant CO₂*WFPS² interaction in the ANCOVA) and moved downward with warming (causing a significant warming effect after adjusting for WFPS effects in the ANCOVA), while the relationship between CH₄ uptake and WFPS was similar between irrigated and non-irrigated plots. These results could indicate that other factors than soil moisture (e.g., changes in soil NH₄⁺, labile C, and/or microbial community composition) were responsible for the shifts caused by elevated CO₂ and warming. However, these shifts are relatively small compared to the large errors associated with each data point in Figure 4. Regardless, our results suggest that soil moisture is the most important factor for explaining the CO₂ and warming effects on CH₄ uptake in this semiarid system.

[22] We have shown that during dry soil conditions, CH₄ uptake in this semiarid grassland responded very differently to elevated CO₂ and warming compared to ecosystems with wetter soil conditions. While CH₄ uptake often decreases in response to elevated CO₂ and increases in response to warming under wetter soil conditions [*Peterjohn et al.*, 1994; *Ineson et al.*, 1998; *Ambus and Robertson*, 1999; *Phillips et al.*, 2001; *McLain et al.*, 2002; *Sjögersten and Wookey*, 2002; *McLain and Ahmann*, 2008], we observed the opposite during times when soil moisture was below the optimum soil moisture content for CH₄ uptake (Figure 6). Indeed, CH₄ uptake responses to elevated CO₂ and warming effects were more sensitive under dry soil conditions. Because the effect of elevated CO₂ on CH₄ uptake was opposite to the warming effect, our results also suggest that combined effects of elevated CO₂ and warming on CH₄ uptake could be less than when only one of these climate change factors is considered. Despite uncertainty about future changes in precipitation [*Christensen et al.*, 2007], recently some have suggested considerable drier conditions for western North America [*Overpeck and Udall*, 2010]. Our results suggest that under drier conditions CH₄ uptake in these grassland ecosystems, which occupy roughly 11% of the global land surface [*Bailey*, 1979], will be more sensitive to elevated CO₂ and warming.

[23] **Acknowledgments.** We thank Elise Pendall, Rebecca Phillips, Dennis Baldocchi, and two anonymous reviewers for helpful comments on a previous version of the manuscript. We thank Erik Hardy, Dan LeCain, Valerie O'Neill, David Smith, Mary Smith, and Katie Tylka for technical assistance. This research was supported by CSREES (2008-35107-18655) and by the Agricultural Research Service under the ARS GRACenet Project.

References

- Ambus, P., and G. P. Robertson (1999), Fluxes of CH₄ and N₂O in aspen stands grown under ambient and twice-ambient CO₂, *Plant Soil*, 209, 1–8, doi:10.1023/A:1004518730970.
- Bailey, H. P. (1979), Semiarid climates: their definition and distribution, in *Agriculture in Semiarid Environments*, edited by A. E. Hall, G. H. Cannell, and H. W. Lawton, pp. 73–97, Springer, New York.

- Bowden, R. D., K. M. Newkirk, and G. M. Rullo (1998), Carbon dioxide and methane fluxes by a forest soil under laboratory-controlled moisture and temperature conditions, *Soil Biol. Biochem.*, *30*, 1591–1597, doi:10.1016/S0038-0717(97)00228-9.
- Castro, M. S., P. A. Steudler, J. M. Melillo, J. D. Aber, and R. D. Bowden (1995), Factors controlling atmospheric methane consumption by temperate forest soils, *Global Biogeochem. Cycles*, *9*(1), 1–10, doi:10.1029/94GB02651.
- Christensen, J. H., et al. (2007), Regional climate projections, in *Climate Change 2007: The Physical Science Basis. Contribution of Working Groups I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change*, edited by S. Solomon et al., Cambridge Univ. Press, Cambridge, U. K.
- Christensen, T. R., A. Michelsen, S. Jonasson, and I. K. Schmidt (1997), Carbon dioxide and methane exchange of a subarctic heath in response to climate change related environmental manipulations, *Oikos*, *79*, 34–44, doi:10.2307/3546087.
- Del Grosso, S. J., et al. (2000), General CH₄ oxidation model and comparisons of CH₄ oxidation in natural and managed systems, *Global Biogeochem. Cycles*, *14*(4), 999–1019, doi:10.1029/1999GB001226.
- Dermody, O., J. Weltzin, E. Engel, P. Allen, and R. Norby (2007), How do elevated [CO₂], warming, and reduced precipitation interact to affect soil moisture and LAI in an old field ecosystem?, *Plant Soil*, *301*, 255–266, doi:10.1007/s11104-007-9443-x.
- Dijkstra, F. A., D. Blumenthal, J. A. Morgan, E. Pendall, Y. Carrillo, and R. F. Follett (2010), Contrasting effects of elevated CO₂ and warming on nitrogen cycling in a semiarid grassland, *New Phytol.*, *187*, 426–437, doi:10.1111/j.1469-8137.2010.03293.x.
- Harte, J., M. S. Tom, C. Fang-Ru, B. Feifarek, A. P. Kinzig, R. Shaw, and K. Shen (1995), Global warming and soil microclimate: Results from a meadow-warming experiment, *Ecol. Appl.*, *5*, 132–150, doi:10.2307/1942058.
- Hutchinson, G. L., and A. R. Mosier (1981), Improved soil cover method for field measurement of nitrous oxide fluxes, *Soil Sci. Soc. Am. J.*, *45*, 311–316, doi:10.2136/sssaj1981.03615995004500020017x.
- Huxman, T. E., et al. (2004), Convergence across biomes to a common rain-use efficiency, *Nature*, *429*, 651–654, doi:10.1038/nature02561.
- Ineson, P., P. A. Coward, and U. A. Hartwig (1998), Soil gas fluxes of N₂O, CH₄ and CO₂ beneath *Lolium perenne* under elevated CO₂: The Swiss free air carbon dioxide enrichment experiment, *Plant Soil*, *198*, 89–95, doi:10.1023/A:1004298309606.
- Kang, H., C. Freeman, and T. W. Ashendon (2001), Effects of elevated CO₂ on fen peat biogeochemistry, *Sci. Total Environ.*, *279*, 45–50, doi:10.1016/S0048-9697(01)00724-0.
- Kettunen, R., S. Saarnio, P. Martikainen, and J. Silvola (2005), Elevated CO₂ concentration and nitrogen fertilisation effects on N₂O and CH₄ fluxes and biomass production of *Phleum pratense* on farmed peat soil, *Soil Biol. Biochem.*, *37*, 739–750, doi:10.1016/j.soilbio.2004.09.010.
- Kimball, B. A., M. M. Conley, S. Wang, X. Lin, C. Luo, J. Morgan, and D. Smith (2008), Infrared heater arrays for warming ecosystem field plots, *Global Change Biol.*, *14*, 309–320, doi:10.1111/j.1365-2486.2007.01486.x.
- Liu, W., Z. Zhang, and S. Wan (2009), Predominant role of water in regulating soil and microbial respiration and their responses to climate change in a semiarid grassland, *Global Change Biol.*, *15*, 184–195, doi:10.1111/j.1365-2486.2008.01728.x.
- McHale, P. J., M. J. Mitchell, F. P. Bowles, and P. P. Bowles (1998), Soil warming in a northern hardwood forest: Trace gas fluxes and leaf litter decomposition, *Can. J. For. Res.*, *28*, 1365–1372, doi:10.1139/cjfr-28-9-1365.
- McLain, J. E. T., and D. M. Ahmann (2008), Increased moisture and methanogenesis contribute to reduced methane oxidation in elevated CO₂ soils, *Biol. Fertil. Soils*, *44*, 623–631, doi:10.1007/s00374-007-0246-2.
- McLain, J. E. T., T. B. Kepler, and D. M. Ahmann (2002), Belowground factors mediating changes in methane consumption in a forest soil under elevated CO₂, *Global Biogeochem. Cycles*, *16*(3), 1050, doi:10.1029/2001GB001439.
- Melillo, J. M., A. D. McGuire, D. W. Kicklighter, B. Moore, C. J. Vorosmarty, and A. L. Schloss (1993), Global climate change and terrestrial net primary production, *Nature*, *363*, 234–240, doi:10.1038/363234a0.
- Miglietta, F., M. R. Hoosbeek, J. Foot, F. Gigon, A. Hassinen, M. Heijmans, A. Peressotti, T. Saarinen, N. van Breemen, and B. Wallén (2001), Spatial and temporal performance of the miniFACE (Free Air CO₂ Enrichment) system on bog ecosystems in northern and central Europe, *Environ. Monit. Assess.*, *66*, 107–127, doi:10.1023/A:1026495830251.
- Morgan, J. A., et al. (2004), Water relations in grassland and desert ecosystems exposed to elevated atmospheric CO₂, *Oecologia*, *140*, 11–25, doi:10.1007/s00442-004-1550-2.
- Mosier, A., D. Schimel, D. Valentine, K. Bronson, and W. Parton (1991), Methane and nitrous oxide fluxes in native, fertilized and cultivated grasslands, *Nature*, *350*, 330–332, doi:10.1038/350330a0.
- Mosier, A. R., J. A. Morgan, J. Y. King, D. LeCain, and D. G. Milchunas (2002), Soil-atmosphere exchange of CH₄, CO₂, NO_x, and N₂O in the Colorado shortgrass steppe under elevated CO₂, *Plant Soil*, *240*, 201–211, doi:10.1023/A:1015783801324.
- Mosier, A. R., W. J. Parton, R. E. Martin, D. W. Valentine, D. S. Ojima, D. S. Schimel, I. C. Burke, E. Carol Adair, and S. J. Del Grosso (2008), Soil-atmosphere exchange of trace gases in the Colorado shortgrass steppe, in *Ecology of the Shortgrass Steppe: A Long-Term Perspective*, edited by W. K. Lauenroth and I. C. Burke, pp. 342–372, Oxford Univ. Press, Oxford, U. K.
- Nelson, J. A., J. A. Morgan, D. R. LeCain, A. Mosier, D. G. Milchunas, and B. A. Parton (2004), Elevated CO₂ increases soil moisture and enhances plant water relations in a long-term field study in semi-arid shortgrass steppe of Colorado, *Plant Soil*, *259*, 169–179, doi:10.1023/B:PLSO.0000020957.83641.62.
- Niklaus, P. A., D. Spinnler, and C. Körner (1998), Soil moisture dynamics of calcareous grassland under elevated CO₂, *Oecologia*, *117*, 201–208, doi:10.1007/s004420050649.
- Overpeck, J., and B. Udall (2010), Climate change: Dry times ahead, *Science*, *328*, 1642–1643, doi:10.1126/science.1186591.
- Peterjohn, W. T., J. M. Melillo, P. A. Steudler, K. M. Newkirk, F. P. Bowles, and J. D. Aber (1994), Responses of trace gas fluxes and N availability to experimentally elevated soil temperatures, *Ecol. Appl.*, *4*, 617–625, doi:10.2307/1941962.
- Phillips, R. L., S. C. Whalen, and W. H. Schlesinger (2001), Influence of atmospheric CO₂ enrichment on methane consumption in a temperate forest soil, *Global Change Biol.*, *7*, 557–563, doi:10.1046/j.1354-1013.2001.00432.x.
- Potts, D. L., T. E. Huxman, J. M. Cable, N. B. English, D. D. Ignace, J. A. Eilts, M. J. Mason, J. F. Weltzin, and D. G. Williams (2006), Antecedent moisture and seasonal precipitation influence the response of canopy-scale carbon and water exchange to rainfall pulses in a semi-arid grassland, *New Phytol.*, *170*, 849–860, doi:10.1111/j.1469-8137.2006.01732.x.
- Rustad, L. E., and I. J. Fernandez (1998), Experimental soil warming effects on CO₂ and CH₄ flux from a low elevation spruce-fir forest soil in Maine, USA, *Global Change Biol.*, *4*, 597–605, doi:10.1046/j.1365-2486.1998.00169.x.
- Sjögersten, S., and P. A. Wookey (2002), Spatio-temporal variability and environmental controls of methane fluxes at the forest-tundra ecotone in the Fennoscandian mountains, *Global Change Biol.*, *8*, 885–894, doi:10.1046/j.1365-2486.2002.00522.x.
- Tom, M. S., and J. Harte (1996), Methane consumption by montane soils: Implications for positive and negative feedback with climatic change, *Biogeochemistry*, *32*, 53–67, doi:10.1007/BF00001532.
- van den Pol-van Dasselaar, A., M. L. van Beusichem, and O. Oenema (1998), Effects of soil moisture content and temperature on methane uptake by grasslands on sandy soils, *Plant Soil*, *204*, 213–222, doi:10.1023/A:1004371309361.
- von Fischer, J. C., G. Butters, P. C. Duchateau, R. J. Thelwell, and R. Siller (2009), In situ measures of methanotroph activity in upland soils: A reaction-diffusion model and field observation of water stress, *J. Geophys. Res.*, *114*, G01015, doi:10.1029/2008JG000731.
- West, A. E., P. D. Brooks, M. C. Fisk, L. K. Smith, E. A. Holland, C. H. Jaeger III, S. Babcock, R. S. Lai, and S. K. Schmidt (1999), Landscape patterns of CH₄ fluxes in an alpine tundra ecosystem, *Biogeochemistry*, *45*, 243–264, doi:10.1007/BF00993002.

F. A. Dijkstra, Faculty of Agriculture, Food and Natural Resources, University of Sydney, Level 4, Biomedical Building - C81, Australian Technology Park, Eveleigh, NSW 2015, Australia. (feike.dijkstra@sydney.edu.au)

R. F. Follett, Soil, Plant, and Nutrient Research Unit, ARS, USDA, Fort Collins, CO 80523, USA.

J. A. Morgan, Rangeland Resources Research Unit, ARS, USDA, Fort Collins, CO 80523, USA.

J. C. von Fischer, Department of Biology, Colorado State University, Fort Collins, CO 80523, USA.